

CONVENTION ON INTERNATIONAL TRADE IN ENDANGERED SPECIES
OF WILD FAUNA AND FLORA



Fourteenth meeting of the Conference of the Parties
The Hague (Netherlands), 3-15 June 2007

CONSIDERATION OF PROPOSALS FOR AMENDMENT OF APPENDICES I AND II

A. PROPOSAL

Inclusion of *Anguilla anguilla* (L). in Appendix II in accordance with Article II §2(a).

Qualifying criteria (Conf. 9.24 (Rev. CoP13) Annex 2a)

A. It is known, or can be inferred or projected, that the regulation of trade in the species is necessary to avoid it becoming eligible for inclusion in Appendix I in the near future.

This species (the European eel) most likely comprises one single stock which is distributed in most coastal waters and freshwater ecosystems in all of Europe, northern Africa and the Mediterranean parts of Asia. For several decades the decline of the stock has been noted. In 2003 an International Eel Symposium provided evidence, based on the four longest glass eel collection series, that the recruitment of young eels to the continental stock had declined to as low as 1- 5 % of its former level in the late 1970s (Fig. 1). The ICES/EIFAC Eel Working Group (2006) analysed the trends of all glass eel collection series up until 2005 and found that the average decline was in the order of 95-99% in the period 1980 and until present. The need for radical management actions was reiterated since eel does not fall under protection of any international law. The scientific community further argued that precautionary action be taken, e.g. by curtailing exploitation and limiting international trade. Export of juvenile eels (glass eels) for aquaculture in Asia (far outside its natural distribution area) comprised more than 50% of the total estimated landing of glass eel since the late 1990s until today. The long and steady decline of this commercially exploited species clearly qualifies it for listing under this criterion.

B. It is known, or can be inferred or projected, that regulation of trade in the species is required to ensure that the harvest of specimens from the wild is not reducing the wild population to a level at which its survival might be threatened by continued harvesting or other influences.

The stock of *Anguilla anguilla* is outside safe biological limits. Most EU Member States acknowledge the serious state of the stock and are concerned about the need of action for recovery within the Community to conserve the stock. The EU Commission has proposed short-term and long-term measures discussed among the Member States. There are still some hesitations and discussions about the Commission's proposal that have to be worked out before it can be adopted by the Council.

High market demand, despite very high market value, has therefore caused opposition to sustainable management proposals in some EU Member States, mainly because different life stages of eel are targeted in several countries. The youngest eel stages (glass eel and elvers) are heavily exploited as they are the basis of eel aquaculture world wide; older eels are also heavily exploited and their migration into and from rivers is impeded by dams and hydropower stations. Even if both current eel fisheries and eel aquaculture in Europe is based on young eel mainly imported from France, Great Britain and Spain and traded within the EU, a substantial part of European glass-eel catches are traded on the Asian market, mainly to China and Japan. Some 90% of eel consumed in the world is based on eel aquaculture, but like direct fishing, this is based on young eel caught in the wild. The glass eel stage is by far the most commercially important life stage and a substantial proportion of European glass-eel catches are traded on Asian markets.

Without trade regulation the species will decline irreversibly both from a commercial and biological standpoint. Therefore a CITES listing of *Anguilla anguilla* is an appropriate measure that would have a beneficial impact on the species.

B. PROPONENT

Germany, on behalf of the European Community Member States acting in the interest of the European Community

(This proposal has been prepared by Sweden).

C. SUPPORTING STATEMENT

1. Taxonomy

1.1 Class: Osteichthyes

1.2 Order: Anguilliformes

1.3 Family: Anguillidae

1.4 Species: *Anguilla anguilla* Linné, 1758

1.5 Common names:	English	eel
	Spanish	anguila, angula (= youngest life stage)
	French	anguille, pibale, civelle
	Swedish	ål
	Danish	ål
	Italian	anguilla
	German	Aal
	Portugese	enguia
	Slovakian	úhor európsky
	Slovenian	

2. Overview

2.1 The European eel occurs in Europe, northern Africa and the Mediterranean parts of Asia. It actually may occur in all ICES fishing areas in the north-east Atlantic except for the areas directly east of Greenland and the Spitsbergen area north of continental Norway (Fig. 2). Within its distribution area it can not be confused with any other species of fish with its elongated snake-like body and smooth slimy skin. Before reaching sexual maturity the eel can reach a length of well over 1 m and a weight of several kilos. It can also attain a very high

age, well over 50 years. The species most likely comprises one single stock spawning in the Sargasso Sea. The eggs hatch there and the larvae drift in a north by north-easterly direction until they reach the European coasts (after 1 - 3 years) and transform through a number of stages to glass eels, elvers, yellow eel and finally into silver eel – the latter being the early sexually maturing stage which seeks to return to the Sargasso Sea to spawn and subsequently die. (Ginneken and Maes 2005)(Maes et al 2006).

2.2 The meat of *Anguilla anguilla* is highly valued in Europe and parts of east Asia. The human consumption preference varies throughout the eel's distribution. In some countries the small, almost transparent glass eels and elvers are highly valued (200 – 1,000 euro/kg), in other countries various yellow eel size groups are sought after and in other countries (mainly in northern Europe) large silver eels, on their way to mature, fetch the best price (5-10 euro/kg). Seen in a global scale the glass eel/elver stage is by far the commercially most important life stage, because almost all "meat production" of eel is based on aquaculture of wild caught young eel stages. The European aquaculture produces half the total supply in Europe while the Asian aquaculture produces almost all Asian supply. Asian eel aquaculture is about tenfold the European production (Dekker 2003a).

2.3 All available information indicates that some types of current European eel fisheries are not sustainable. Recruitment has been declining since the 1980s and reached a historical low in 2001 and has not improved since then. Eels are exploited in all life stages and fishing mortality is high. In addition to overfishing, other anthropogenic factors might have contributed to the sharp population decline: inland (freshwater) and coastal habitat loss, pollution, climate change, ocean current change and loss of upstream / downstream migration routes through for example hydroelectric power stations and other constructions.

2.4 According to ICES/EIFAC Working Group on Eels and ICES Advisory Committee on Fishery Management (ACFM) a recovery plan is urgently needed and the European Commission has requested ICES to evaluate what mitigative measures should be instituted to improve the situation. Considering the many uncertainties in eel management and the uniqueness of the single eel stock a precautionary reference point for eel must be stricter than universal provisional reference points. Exploitation should be reduced to as close to zero as possible until such a recovery plan is agreed on and implemented.

2.5 National monitoring of the various eel stages is fragmentary. Some traps on rivers provide fairly reliable data on upstream migration of young yellow eels, but there are virtually no regular routine surveys of yellow and silver eel in fresh water or along the coasts. Some of the long-term series may also be terminated in the near future as a consequence of decreased turnover of local fisheries and the impossibility of addressing this large-scale stock decline at the local level. There are also inconsistencies between official statistics on eel landings and ICES estimates. A major revision of data bases is thus also required.

2.6 *Anguilla anguilla* meets the guidelines suggested by FAO for the listing of commercially exploited aquatic species. The species actually falls into FAO's lowest productivity category of the most vulnerable species and the rate of decline is so rapid and steep as to qualify for Appendix I listing under these FAO guidelines. The latest IUCN Red List assessment for this species is (probably) that of Sweden (2005) which lists the European eel as *Critically Endangered* (CR).

2.7 An Appendix II (App. B) listing for *Anguilla anguilla* will regulate and monitor future international trade, hopefully ensuring that future fisheries will not be detrimental to the status of the wild stock and thus to the survival of the species. This legal measure will also facilitate traditional eel management measures, and the Community-level measures for a coordinated recovery plan that is currently being developed by the European Commission.

2.8 Even though research in Japan has come quite far, artificial reproduction is still not possible for European eel, all aquaculture and restocking is still based on capture of wild young eel. Even though there is some concern that disease and reduced genetic variability may result from restocking, this risk must be balanced against the potential benefit from this measure, and the risk of further stock decline due to a failure to take this action. The most recent research has shown that the European eel is still believed to be panmictic and that the genetic variation found is mainly a temporal one (between batches/cohorts within year) and not a spatial variation. (Albert et al. 2006, Danevitz et al. 2005., Maes et al. 2006 a, b., Pujolar et al. 2006) Thus genetics are of minor concern. See also below under 3.1.

It has been estimated that present catches of glass eel in Europe cover only some 1/6 of the demand of the European market for re-stocking, not counting the aquaculture demand in Asia and Europe!

3. Species characteristics

3.1 Distribution

The European eel (*Anguilla anguilla*) occurs from the Atlantic coast of North Africa, in all of Europe, including the Baltic Sea and in the Mediterranean waters of Europe, northern Africa and Asia. In addition the European eel also occurs in the Canary Islands, Madeira, the Azores and in Iceland (Schmidt 1909). The latter island is probably unique because it also harbours American eels (*Anguilla rostrata*). Furthermore, there is also evidence of interbreeding of the two eel species occurring there (Avisé et al. 1990). It is important to realise that the European eel is believed to spawn in the eastern part of the Sargasso Sea (although spawning has never been directly observed) so the distribution of eels on their spawning migration extends all the way from northern Europe across the Atlantic Ocean and down to the Sargasso Sea, north by north-east of the West Indies. The newly hatched larvae drift with the Gulf Stream and the North Atlantic Current to the continental shelf of Europe and North Africa thus closing the life history distribution of the European eel. It has been generally accepted that the European eel comprises a single panmictic stock (e.g. Schmidt 1925, DeLigny and Pantelouris 1973, Tesch 1977, Avisé, Helfman, Saunders and Hales 1986, Lintas, Hirano and Archer 1998). A recent study (Wirth and Bernatchez 2001) using highly polymorphic gene markers provided evidence of genetic differentiation. These authors found that the distribution of genotypes were indicative of non-random mating and indeed of restricted gene flow among eels from the three broad groups found – the Mediterranean, the North Sea and Baltic and the northern groups (Iceland) respectively. These findings of course would have far-reaching implications for eel management. However, more recent studies (Dannevitz et al 2005., Albert et al 2006., Pujolar et al 2006., Maes et al 2006 a,b.) indicates a more subtle, temporal pattern, that might have appeared as a spatial pattern in the study of Wirth and Bernatchez, due to unsynchronised sampling in northern and southern areas. However, even though the exact identity of the Icelandic stock might be disputed, the abundance of eel in Iceland is that low, that neither fishing nor trading of eels from Iceland plays any role at the population level (Dekker 2003b). Whether a single panmictic stock or a species with as more complex stock structure, the management of the European eel must be co-ordinated to ensure adequate escapement throughout the species range (Russel and Potter 2003).

3.2 Habitat

Although the European eel is considered a temperate species it also occurs as spawning adults and newly hatched larvae in the *tropical waters* of the Sargasso Sea, in the *sub-tropical waters* of the Azores, the Canary Islands, Madeira, the Atlantic coast of north-western Africa and the African coast of the Mediterranean, and, in the frigid *arctic waters* of Iceland, Jan Mayen and northernmost Norway (Schmidt 1909). However, the high yield of eel production and fisheries in temperate areas is in contrast with the temperature preference of the species which ranges from 10-38 degrees centigrade, with an optimum around 22-23 degrees (Boetius and Boetius 1967, Sadler 1979, Dekker 2003b).

The northern distribution area has no sharp limit, the density of eels simply gradually fades out (Dekker 2003). The conventional view is that eels are catadromous, i.e. they spawn in salt (marine) habitat and then move into freshwater areas to grow as yellow eels and subsequently become sexually mature (silver eels) (Table 1). However, yellow eels can also be found in estuarine and coastal habitats throughout the area where glass eels and elvers occur naturally, and some may actually remain in marine habitat their entire life-cycle (Tsukamoto, Nakai and Tesch 1998, Daverat et al. 2006). For the Baltic Sea is noted that around 80% of the eel remain in this marine habitat for all their life (Wickström and Westerberg 2006).

In summary, the European eel occurs in an extremely variable number of habitats during its life cycle: 1) spawning, newly hatched larvae and all marine developmental stages occur in the marine pelagic zone of the Atlantic Ocean, 2) glass eels, elvers, some yellow eels and some silver eels occur throughout their life in shallow marine coastal areas, 3) some glass eels, elvers, yellow eels and silver eels move into and/or grow in coastal lagoons and estuaries, 4) some glass eels, elvers, yellow eels and silver eels move into or grow in freshwater habitats, swim upstream brooks and rivers and further into ponds, lakes and reservoirs, where they may remain for decades before they ultimately swim downstream on their final spawning migration. Clearly, any habitat destruction occurring in any type of water body will negatively affect the European eel.

3.3 Biological characteristics

Some basic biological characteristics of the eel have been described above. Suffice it to repeat that most researchers still agree with the views of Schmidt (1909, 1925) that the European eel comprise a single panmictic stock which spawns in the Sargasso Sea. Although spawning has never been observed newly hatched larvae have been observed from a relatively small area of the Sargasso Sea (Schmidt 1922). Schmidt also followed the increase in size of the various developmental stages of these larvae (leptocephali) and thus could map their migration (actually drift with the currents) to the north-west African and west European coasts. The leptocephalus larvae metamorphose into glass eels when they arrive at the continental shelves of north-western Africa and Europe after a journey of approx. 3 years (Tesch 2003).

Eventually, the glass eels become pigmented elvers as they either enter estuaries, brooks and rivers where they spend their growth phase as yellow eels, or actually spend their entire growth phase in either brackish or marine habitat (Daverat et al. 2006). The growth phase may last from 3 to up to at least 25 years, depending on sex and environmental conditions. On average males migrate at an age of 7-8 years and females at approx. 11 years of age (Tesch 1977). A female eel may attain a weight of more than 6 kg and a length of well over 1 meter whereas the males rarely exceed 45 cm in length (Wickström 2005). Eels are also long-lived fishes. In captivity an eel was recorded to have lived for 84 years. At the start of migration the gonads gradually mature and the eels migrate back to the Sargasso Sea to spawn and die.

There is no evidence of any eel surviving spawning. A good overview of the life cycle and the major life stages of the European eel is given by Dekker (2000a). See also Fig. 3.

The gonads of eels are undifferentiated until a length of 15-25 cm (Kuhlmann 1975). In natural waters this size is attained in the yellow eel stage a few years after the glass eel stage (Tesch 1977). Some authors argue that sex differentiation is environmentally influenced (e.g. Parsons et al. 1977, Wiberg 1983) whereas others claim that different migratory behaviour of females and males account for the difference (D'Ancona 1958, Svårdson 1976). Holmgren (1996) in her doctoral thesis on sex differentiation and growth pattern in the European eel concludes that her results show that females may develop in any habitat type, but males should only develop if they experience good conditions for growth during the early gonad differentiation, which may be independent of the resources needed for growing to a large silver eel size. Yellow eels that migrate far up in river systems have probably not met this criterion and will thereby become females. This information is given because it has both management implications as well as economic significance. Eel farmers want to optimise early weight increase and will consequently favour male eels. On the other hand may young eels that have been stocked in natural lakes develop in either direction, depending on individual growth performance, before or after they enter the new environment (Holmgren 1996).

3.4 Morphological characteristics

The European eel is one of approx. 15 anguillid eel species in the world. They all resemble one another by being long, slender and snake like with almost cylindrical bodies covered with very small scales. Their skin is smooth and slimy. Eels lack ventral fins and the dorsal-, tail- and anal fins form a continuous fin from the mid section of the back to the anal opening. The gill openings are small. As described above the eel passes through a series of developmental stages during its life cycle: 1) the transparent leptocephalus marine stages, 2) the more cylindrical but still transparent glass eel, 3) the pigmented elver and then through 4) the long yellow eel period of the growing eel to 5) the migrating silver eel which has ceased to feed and spends its energy resources entirely on production of gonads and the long migration back to the Sargasso Sea. During the silver eel stage the eel changes its colour and appearance considerably. Whereas the "yellow eel" is grey/green/olive/brown on the back side and yellow/green/white on the ventral side, the silver eels turn into a more "marine appearance" with dark and even black back and a silvery or copper coloured ventral side. In addition the eyes become larger and the lateral line more pronounced (Wickström 2005). These differences between the yellow eel and silver eel stages occurs in both sexes.

3.5 Role of the species in its ecosystem

The role of the eel in its ecosystems is a many-faceted issue, because, as argued above, the eel belongs to so many different ecosystems during its life cycle. The marine larval stages of eel probably feed on microscopic plankton and remains of plankton and very likely have no effect on the pelagic ecosystem in which they live for approx. 3 years. During the glass eel and elver stages probably larger prey may be taken because these stages are far more mobile than the younger ones. Finally during the yellow eel stage – the growth period – eels are opportunistic omnivorous predators. Chironomid larvae, worms, mussels, gastropods, insects, crustaceans (freshwater crayfish in particular), fishes and fish roe are consumed when available, even frogs and small rodents may be eaten. The only instance when a conspicuous effect of yellow eels on their ecosystem has been noted is when freshwater crayfish (*Astacus astacus*) have been present in the river or lake. After eels have been stocked some crayfish populations have been severely depleted by eel predation. Otherwise eels do not seem to

affect significantly the recruitment of other species. This broad diet would indicate that eels were quite susceptible to other predators, but contrary to this hypothesis yellow eels show very high survival rates. Moriarty (1987) attributes this success of eels to avoidance of all predators (at all life stages), and also on high survival during sub-optimal conditions for growth. Even when glass eels are stocked into lakes where this eel stage would never occur naturally a very large percentage have managed to survive until they have been recaptured as yellow or silver eels (Tulonen and Pursiainen 1992).

4. Status and trends

4.1 Habitat trends

The environmental threats to eel habitat include barriers to upstream migration but also hydroelectric facilities where the turbines may seriously impair the downstream migration of silver eels causing high mortality. Eel ladders and bypasses may on the other hand mitigate both hazards. Another factor impairing the reproductive capacity of the eel is bio-accumulation of lipophilic contaminants and concentration levels in the fat of their muscles and gonads seems to be a reflection of the actual concentrations in the environment (ICES 2006). In general, due to the high energy costs of the spawning migration the adipose tissue energy stores are gradually depleted and the contaminants found in the adipose tissue may impair the success of reproduction. Pollution of the benthos is thus a threat to the yellow eel stage. Extensive and unregulated live transport of eel of all sizes is another potential danger, because parasites and viruses can spread both to wild populations and to dense populations in aquaculture. There is no general trend in a favourable direction for eel habitat but reduced emissions of some toxins will have beneficial effects in future, as will construction of fish ladders, bypasses and better grids at hydropower stations and other obstructions to eel migration.

4.2 Population size

4.2.1 Spawning stock

As mentioned above the natural spawning behaviour of the eel has never been observed directly nor do we exactly know the exact location, timing and abundance of eels in the spawning area. In addition, sampling methods have not been standardised (Moriarty and Dekker 1997) so comparison of stock density among catchments and countries is rarely appropriate. Despite this serious lack of knowledge management measures must be enforced to protect the spawning stock regardless of time, place and size. This management advice follows the precautionary approach - PA (ICES 1999). The management targets aim at protection and recovery of the spawning stock. In accordance with the PA measures should aim at protecting 30% of pristine spawner escapement and an extra safety margin has been recommended (ibid.) to protect 50% of this escapement.

Dekker (2000b) noted that the number of silver eels escaping to the ocean on their spawning migration is negligible in comparison with commercial landings. As a consequence, variation in yellow eel fishing intensity will cause the mean age in the catch to vary, but will only affect the number of eels caught marginally (Dekker 2003b). Obviously, the commercial eel catch provides an index of stock size.

Estimation of the potential spawning stock should rely on historical data (Dekker 2003a). Because only information on recruitment is available estimation of the spawning stock must be based on modelling of population dynamics. Models of the continental phase of eel population dynamics have been developed following three lines: 1) the Leslie-matrix cohort-model approach (Gatto and Rossi 1979), 2) the Input-Output approach which directly relates

juvenile recruitment abundance to migrating silver eels (Völlestad and Jonsson 1988), and 3) a number of models ranging from stage-structure and density dependence survival from one stage to the next to more complex size/age/stage structured models (e.g. De Leo and Gatto 1995, Dekker 1996, Reid 2001, Greco et al. 2003, Åström 2005).

These models of course differ in terms of mathematical complexity and usability. While site-specific analyses are needed to frame the eel life history in the continental phase, the generalised decline of eel recruitment requires a global assessment of meta-population viability.

The first attempt to calculate the size of the European eel stock was performed by Dekker (2000b; diagram in Dekker 2003a; Fig.4 here). Dekker also calculated the dynamics of the eel population in the early 1990s. Further research in this area is ongoing and will help to improve estimates of stock abundance both in the past and present situations (ICES 2006). It is hoped that these models can be adapted also to areas where little data is available.

4.2.2 Panmixia, recruitment and production

As mentioned above most eel biologists argue that the European eel comprises a single panmictic stock. Even though we know that this eel species is wide-spread and in drastic decline, available data on recruitment, stock and fisheries are still fragmentary. Obviously, almost all water bodies within its natural distribution contain, or have contained, eels in a few or all pigmented stages. This means that the eel population is fragmented into thousands of water bodies. Already in 1997 Moriarty and Dekker noted that “recruitment has steadily decreased since the early 1980s, fisheries have declined and man-made impacts on the habitats of this species have adversely affected production potentials.” A few years later Dekker (2000) argued that the absence of sufficient data on the myriad of small local sub-stocks preclude a reliable stock assessment. However scanty the data on total population size compilation of FAO data-bases in the 1990s indicated that the world-wide production of anguillid species in fisheries in the order of 30,000 tonnes per year. Roughly one half of that catch comprised the European eel (Dekker 2003a). In order to improve assessment of the biological status of the eel, this species has been included in the EU Data Collection Regulation, but required sampling levels have only been tentatively indicated, however, only a few countries have included eels in the national sampling programmes.

Contrary to common belief it seems likely that more than 60% of eel production takes place in coastal marine habitats (Wickström and Westerberg 2006). Actually, some 80% of all eels leaving the Baltic have spent their entire life in salt water habitats. These authors (ibid.) also argue that this proportion may increase with declining recruitment. Consequently, it is necessary to include all marine eel fisheries in an European Eel Management Plan.

4.3 Population structure

As described above this species is highly migratory and comprises a series of developmental stages throughout its life cycle, which tends to segregate the species geographically by age. As a consequence, different nations within the eel’s distribution have developed fisheries which may target different age stages, actually covering both glass eels, elvers, yellow eels and silver eels. As a result it is unlikely that a natural population structure exists in the various regions where there is a fishery for the different life stages.

4.4 Population trends and geographic trends

4.4.1 General trends

The generation time for *Anguilla anguilla* defined as the average reproductive age of females varies between sub-populations but is approx. 11 years, in some northern sub-populations often 15-20 years and even older. The three-generation period against which declines must be assessed (Annex 5, CoP9.245, Rev. CoP13) is thus some 30-35 years upwards to 60 years.

Few data sets provide information on changes in the level of recruitment and those that are available relate to various stages of the recruitment into continental habitats (Dekker 2002). Time series from 19 rivers in 12 countries have been examined for trends. Data from eleven of those rivers are available for 2005 (ICES 2006, Table 2). National trends in glass eel, elver and “young eel” recruitment are shown in Fig. 5. The most conspicuous trend can be seen in the Norwegian River Imsa, where there is no fishery and no stocking, yet a drastic decline in elver recruitment.

Conspicuous downward trends occur in all time series in the last two and a half decades. This is a reflection of the rapid decrease after the 1970s (ibid. Fig. 6). Data collected in the last few years indicate that recruitment now (2006) is even lower than the minimum level of 2001. The low level of recruitment of 2001 was also synchronous with a smaller size of glass eels, which interpreted as a sign of adversory oceanic conditions. The most recent low recruitment levels, however, occurred under more favourable oceanic conditions (NAO index), and mean glass eel length was not lowered. This indicates that most recent low recruitment figures are very unlikely to be caused by adversory oceanic effects (ICES 2006). If the current trend continues, the stock might reach the brink of extinction within a single generation (<10 years). (Dekker 2004) In October 2005, the EC proposed a “Council Regulation establishing measures for the recovery of the stock of European eel” (COM 2005, 472 final).

In northern areas no glass eels are found recruiting the river sub-populations because there the transition to the yellow eel stage happens long before they enter fresh-water habitats. Long-term data series from four northern rivers (1 Norwegian, 3 Swedish) are shown in Figure 7 (ibid.). In the first half of the 1990s a moderate recovery in glass eel recruitment was observed, which later in that decade can be seen as an increase in yellow eel recruitment.

4.4.2 Trends in re-stocking

Data on re-stocking are available from a number of countries. Glass eels and young yellow eels are reported separately. The yellow eel component varies in size (age) among countries and data are presented on a weight basis which then can be converted to numbers, using estimates of average individual weights of re-stocked eels. As an indication of the size variation obtained Denmark reports 3.5 g, 20 for Germany, 33 for the Netherlands and 90g for Sweden. An overview of the trends is shown in Fig. 8 (ibid.).

In European countries other than those combined in those Figures the following information can be given:

Latvia – during Soviet time, starting in the 1960s, roughly 30 million glass eels were stocked into 51 lakes. At present, only a few lakes are stocked and with a low number of glass eels.

Lithuania – re-stocking commenced already in the late 1920s. Since the 1960s some 50 million elvers and young yellow eels have been stocked.

Germany – no central data base for re-stocking but considerable local re-stocking

Ireland – elvers are stocked in some drainages.

France – no central data base for re-stocking.

Spain – no central data base for re-stocking.

Italy – no central data base for re-stocking, but considerable local re-stocking.

4.4.3 Formal status of the eel stock

IUCN (International Union for the Conservation of Nature) has compiled criteria for ranking species in terms of risk of extinction. IUCN recommends that the English abbreviations of the so-called Red List Categories be used irrespective of the language used in compiling the national Red Lists. This practice makes the Red Lists of different countries easier to understand and also help comparison of the status of a species among different countries. The level observed since 1990 is below 20% of the level observed not more than three generations ago. The European eel therefore qualifies for the IUCN Red List of endangered species.

Opportunities for protection and restoration of spawner escapement are fading.

The most extreme categories refer to the fact that a species is totally or regionally extinct. The second most severe condition is when a species is Critically Endangered (CR). The criteria state that at least 2% of the total population resides within a country, and that its stock has declined by 80% or more over not more than 3 generations. Since these criteria are met (see above) Sweden has listed the eel on its national Red List as Critically Endangered (CR). So far, no other country has done so (ICES 2006).

5. Threats

As argued about concerning the trends in population size, recruitment and habitat quality, one of the major threat to this species is over-exploitation by some types of fisheries targeting the various life stages. In addition, blocking of rivers by dams, pollution of waters and sediments and habitat alterations have adverse effect on recruitment and survival of the species as well as introduced parasites such as *Anguillicola crassus*, which may impair the migration of adults. All mentioned reasons have been implicated in reducing spawner quality. Hydropower and drainage pumps together with fisheries are major causes of eel mortality (of the lethal threats for downstream migrating eels). Furthermore, the parallel declines in European and American eels, both of which spawn in the Sargasso Sea, has been taken as evidence that changes in ocean currents resulting from climate change may have interfered with larval transport leading to reduced recruitment in both stocks. While this does not negate the need to reduce the mortality in some types of fisheries, it acknowledges the reality that restricting trade alone may not be sufficient to bring about recovery.

Two principal positive characteristics of the fish and the fisheries is 1) the natural very high survival rate of yellow eels in their various habitats, and 2) the fact that there are hardly any by-catches of eels in gear other than those targeting the species.

5.1 Directed fisheries

Cultural patterns in fishing, aquaculture and consumption determine much of the distribution of various fisheries methods. This is particularly true of glass eel exploitation which interferes with the relation between stock density and fishing yield. In addition, in the 20th century consumption patterns changed dramatically. In the early half of that century glass eel were consumed in England, Wales and Ireland, a tradition entirely lost today. A similar change occurred in France where glass eel were consumed locally, but are now exported to Spain and east Asia (Dekker 2003a). In north European countries, glass eel are caught and used for restocking rather than immediate consumption. In general, fisheries tend to adapt to stock abundance and market options rather cultural traditions.

As mentioned in the Overview, fisheries are directed to various stages of eel in different countries and regions, not only because of local food habits but also because of market prices and demands from the expanding aquaculture industry in Asia and elsewhere. In general,

though, eels are important throughout Europe for the small-scale coastal fisheries (Fig. 9). This also applies to the freshwater fisheries, mainly in northern Europe. Even though the fisheries are small scale and local the market is becoming increasingly global and eel trade is substantial (Wickström 2006). It was stated above that fisheries methods targeting eel rarely yield high by-catches of other fish species. On the other hand, seals and birds may drown in eel-catching gear.

According to FAO data bases it was estimated that the entire catch of eels in Europe was approx. 5,000 tons in 2002. Unofficial sources, however, argue that catches of 30,000 tons annually were caught in the 1990s, a figure that by now may have declined to some 10,000 tons (Wickström 2006). This is in agreement with Moriarty and Dekker (1997) who propose that the annual European catch in the 1990s was some 20,000 tons, Moriarty and Dekker (ibid.) also state that more than 25,000 people in Europe acquire a substantial income from eel fisheries. A comparison of the change in eel catches in Europe between 1994 and 2004 is shown in Table 3.

The fishing yield of European eel amounts to more than half of the world eel fisheries on all eel species. Annual averages in the 1990s, according to FAO data bases, were of the order of approx. 15,000 tons out of a world fisheries catch of some 29,000 tons. The annual average aquaculture production of eel in the 1990s was approx. 208,000 tons, more than 90% of which were the “Japanese eel” (*Anguilla japonica*). In 2002 that figure had increased to more than 230,000 tons according to FAO data bases (160,000 tons by China alone). Also in Europe aquaculture production exceeds fishing yield (Table 4) with three countries accounting for the bulk of production (Fig. 10). All in all, aquaculture production accounts for some 90% of present eel production world wide. Obviously, the fishing of glass eel and elvers provide the bulk of aquaculture production. Commercial glass-eel fisheries are found from the south-western end of the distribution area to River Severn in the north and including the Mediterranean coasts of Spain and Italy. Outside of this area glass eels are also caught but mainly for re-stocking inland waters either to supplement natural eel production or to use traditional growing areas where eels no more ascend the rivers.

Glass-eel fisheries are, as mentioned above, very species specific and no by-catches are obtained. The fishing methods used include hand-held or ship-based nets, either fixed or being moved. A wide range of dipnet types are used, but also trawls, stow nets, and fykenets (e.g. Dekker 2002, Aubrun 1986, 1987, Weber 1986, Ciccotti et al. 2000).

Data from the mid 1990s (Moriarty and Dekker 1997, Dekker 2000b) are presented in Dekker (2003) to show the “use” of glass eels arriving to the European continent and its surrounding waters (Fig. 11). When converting the numbers given in Dekker’s (ibid.) diagram to percentages, the following picture emerges: 50% goes to aquaculture (43% to Asia and 7% to EU countries – mainly Italy), 18% is used for direct consumption (almost all by Spain), 10% is used for trap & transport within EU countries, 8% is traded for re-stocking between countries, and, finally, only 14% escapes as natural immigration.

Yellow and silver eels fisheries are found throughout Europe (Fig. 12). In central and northern Europe these life stages dominate the catches. Even if the glass-eel catches are marginal in weight they outnumber the yellow and silver eels catches by a factor of 30 (Dekker 2000). Downstream migrating silver eels have been fished for hundreds of years in central and northern Europe in fixed traps, both on small streams and in big rivers, but such directed fisheries have dwindled down all over the original area. However, the silver eel

fisheries still dominate the fisheries in Scandinavia. The low density production of yellow eel in northern countries is turned to a highly profitable fishery on the silver eel stage, because they tend to concentrate their emigration both in time and space along the coast. At intermediate densities in central Europe fisheries focus on the yellow eels stage with a “by-catch” of silver eels.

Fisheries on the yellow and silver eel stages apply a wide series of gear: fixed traps (fish houses), all kinds of nets, spears, pots, hooks (longlines) and fyke-nets (e.g. Gabriel 1999).

5.2 Incidental fisheries

The early life stages of the European eel are rarely caught as a by-catch in gear targeting other species of fish. Yellow eel on the other hand are sometimes caught on bottom-set long line hooks baited with worms or small fish. Both yellow and silver eels are also occasionally caught in small-meshed fyke-nets that are non-selective in terms of species of fish caught. Additionally, infrequent by-catches occur in marine bottom gears, such as otter trawls and beam trawls, but these by-catches largely remain unregistered. There is no data available on the percentage of the total catch that these by-catches account for, but an educated guess is that it is a marginal proportion.

6. Utilisation and trade

European eel are utilised as a highly valued human delicacy in most European countries. Some countries mainly consume the glass eel stages, others eat small yellow eels, and still other countries eat the large yellow eels or only silver eels. The international trade of *Anguilla* spp. is high and from Europe the the main export of glass eel is going to Asia (Eurostat data, Table 5) The international trade in *Anguilla* spp., including internal EU transactions, is recorded under four specific codes of the CN (Combined Nomenclature) of the EU and HS (Harmonised System) of custom services around the world. The four specific codes are:

- 0301 92 00, Live eels – “*Anguilla* spp.”
- 0303 66 00, Fresh or chilled eels “*Anguilla* spp.”
- 0303 76 00, Frozen eels “*Anguilla* spp.”
- 0305 49 50, Eels “*Anguilla* spp.”, smoked, incl. fillets

Compared to other fish taxa (e.g. *Lamna nasus*, Porbeagle shark) or most wildlife commodities in trade, this should allow for a quite accurate recording of volumes and values. However, concerning live glass eel trade, data occur very fragmentary and unreliable because current customs nomenclature, make it difficult to distinguish between different species and life stages. In other words, records of live glass eel exports and imports are mixed with the trade in live juvenile and adult eels.

According to Eurostat, trade from Europe to Asia is almost entirely based on glass eel and used in aquaculture and it also shows that the highest price is paid by the Asian countries (Eurostat data, Table 6, Figure 16). The glass eel has the highest value per kilo in trading as the glass eel is very important for aquaculture. The trends of prices paid for glass eels may illustrate the high commercial pressure that national and international trade can have on wild populations of eels around the world. In 1996 for instance, their increasing scarcity associated to the constant demand from Japanese eel farmer, led the retail price of Japanese glass eels to peak up to USD11 800/kg (about 2.4 each) (Ringuet *et al.*, 2002). The “import” value (wholesale) of European glass eels almost reached EUR 200/kg in the late 1990s and rose to EUR 340/kg in the early 2000s (Eurostat data, Table 6 and Figure 16). To note is that the retail price that Japanese eel farmers were ready to pay for glass eels peaked at about EUR

10,000/kg in 1996. Compare this price with the Beluga caviar, considered one of the most valuable fishery commodity, worth about EUR 8000/kg at retail outlets. Based on Table 6 and associated analysis in Ringuet *et al.*, 2002, the estimated total value of the world trade in European glass eels in 1997 (100 to 130 tonnes) was EUR 30 millions (C. Raymakers, pers. comm. November 2006). All this signals the threat to the *Anguilla* spp. represented by the extraordinary commercial pressure on live glass eels.

The Community is an exporting country of eel of all life stages, with prices reaching more than 1000 €/kg in glass eel. Imports of live eel into the Community had an average price of 7.7 €/Kg in 2005, while exports of live eel from the Community had an average price of 704.95 €/Kg, almost 100 times more. (See separate Annex “EU's EXPORT of LIVE EELS, *Anguilla* spp., CN-code 0301 92 00, based on Eurostat data). The reason for this is that import consists mainly of adult eel, while exports consist of glass eel.

According to Dekker (2003a) aquaculture production of European eel (which is based entirely on wild caught glass eel) exceeds the fishing yield of the species. In addition, an unknown amount of glass eel/elvers caught in Europe is exported to East Asia (mainly China and Hong Kong). The rise of aquaculture in Asia and in Europe has initiated a world wide trading web. It is obvious that this, in volume a very small scaled fishery, is in fact a large world wide trade where the Asian demand determines the European prices. Comparing the value and the weight of internal and external exports of eel from EU shows that the price rises although the amount exported is decreasing (Eurostat data, Table 6 and Figure 16). Bad recruitment of *Anguilla japonica* provoked the rises of the value of the European glass eel fishing while the demand for European glass eels decreased in the years when recruitment of *Anguilla japonica* was higher (Ringuet *et al.*, 2002).

From 1995 to 2005, the average number of European glass eels exported annually from the EU was estimated at about half a billion fish, the estimated number is based on the conversion rates from weight to number of individuals (around 3000 glass eel/kg) (Ringuet *et al.* 2002 Table 7). Take into consideration that in restocking of glass eel, within the natural distribution area, around 5-10% of the eels will reach the reproductive stage of a silver eel. Also, elsewhere in this report it was stated that the present number of glass eel caught in European waters is insufficient for European restocking needs, not to mention European and Asian aquaculture demands. Small yellow eels (for stocking purposes) are also traded among European countries and also within countries. This measure is mainly to supplement inland commercial fisheries focusing on silver eels and sometimes also on large yellow eels. These figures (See separate Annex “EU's EXPORT of LIVE EELS *Anguilla* spp., CN-code 0301 92 00, based on Eurostat data) show that trade in eels and in particular export of glass eels to Asian countries has become a very lucrative business providing unquestionably a strong incentive for the harvest of this species.

A CITES Appendix II listing for *Anguilla anguilla* will regulate and monitor future international trade, particularly from Europe to Asia, hopefully ensuring that future fisheries will not be detrimental to the status of the wild stock and thus to the survival of the species. This legal measure will also complement (and reinforce) traditional eel management measures, and the internationally coordinated recovery plan that is currently being developed by the European Commission.

7. Legal instruments

Catadromous species (spawning in the sea but often growing and maturing in inland waters) like the European eel have special attention in international law. United Nations Convention on the Law of the Sea (UNCLOS) has a special Article (67) covering general principles for management of these species. In short, the following rules apply:

- 1) Coastal states/countries are responsible for management, but also states through the territory of which the species migrate are responsible for binding agreements concerning management measures.
- 2) Fishing at sea is allowed within the Exclusive Economic Zone but prohibited in the high seas.
- 3) Management must include provisions for secured immigration and emigration of the species.

These measures at least point at the need for international co-operation in eel management. One such concrete environmental aspect is to make sure that rivers will not cause obstruction to natural eel migration, e.g. because of pollution and construction. Most natural migratory routes to inland waters are now within EU jurisdiction, but some part of the drainages will also affect third countries. These facts call for management to be co-ordinated by multilateral agencies like The European Inland Fisheries Advisory Commission (EIFAC), instituted already in 1957 by FAO, and ICES when scientific advice is warranted.

8.1 Management measures

At the 92st Statutory Meeting of ICES (2005) and at the 25th meeting of EIFAC (2005) it was decided that the **ICES/EIFAC Working Group on Eels (WGEEL)** would meet in January 2006. Main Recommendations by the EIFAC/ICES Working Group on Eels are as follow:

- a) the rapid development and implementation of management plans is facilitated in a work programme of workshops and guidelines, i.a. for
 - re-stocking practices,
 - recruiting eel immigration passages,
 - silver eel deflection schemes,
 - monitoring and post-evaluation procedures, potentially in pilot projects,
 - pollution and disease monitoring,
 - development of models and tools for management of the stock;
- b) areas producing high quality spawners (large sized females, low contaminant and parasite burdens, unimpacted by hydropower stations) be identified in order to maximise protection for these areas;
- c) management targets are set for spawner escapement with reference to the 1950s-1970s, either identifying the actual spawner escapement levels of that period in full, or 30-50% of the calculated spawner escapement that would have existed if no anthropogenic mortalities would have impacted the stock - and where adequate data are absent, with reference to similar river systems (ecology, hydrography);
- d) under the implementation of the WFD eel specific extensions should be implemented as an indicator of river connectivity and ecological and chemical status.

8.1.1 Objective of recovery

The objective of recovery of the stock necessitates restoration of the spawning stock, for which the EC has proposed a target of 40% of the potential production under unfished, unpolluted and unobstructed conditions. A methodology for elaboration of this reference level is described in this report (WGEEL 2006), but actual implementation will require field data and analysis for each spatial management unit. Analysis of stock dynamics under different fisheries management regimes indicates that *recovery times may vary from 20 up to 200 years*, depending on the intensity of implemented fisheries restrictions. However, restrictions on fisheries alone will be insufficient, and management measures aimed at other anthropogenic impacts on habitat quality, quantity and accessibility will also be required (WGEEL 2006). Also, the development of national and international management plans will involve aspects related to the Common Fisheries Policy (CFP) as well as to the Water Framework Directive (WFD). The overall objective will have to be achieved by implementation of protective measures at a regional scale, presumably at the level of River Basin Districts (RBDs) as defined for the Water Framework Directive (WFD).

The latest report (WGEEL 2006) constitutes just one step in an ongoing process of documenting the status of the European eel stock, fisheries and compiling management advice. As such, the Report does not present a comprehensive overview, but should be read in conjunction with previous reports (ICES, 2000; 2002; 2003; 2004; 2005a)

8.1.2 The latest Official Regulation proposed by the EC

The objective of this proposal is to achieve a recovery of the stock of European eel and to ensure the sustainable use (fishing) of the stock. The principal element of the proposed Regulation is the establishment of eel management plans for each River Basin, including trans-boundary basins (as defined according to the Water Framework Directive). The objective of each River Basin management plan shall be to permit, “with high probability, the escapement to sea of at least 40% of the biomass of adult silver eel relative to the best estimate of the potential escapement in the absence of human activities affecting the fishing area or the stock”.

According to the proposal, Management Plans should have been communicated to the Commission by 31 December 2006, and plans approved by the STECF must then be put in place by 1 July 2007. Subsequent monitoring of the effectiveness and outcome of the plan should be communicated to the Commission by 31 December 2009. This proposal has been amended and approved unanimously by the European Parliament, and is now awaiting discussion by the Council of Ministers. Schedule of the process can still be changed depending on future discussions. The EU web site describing the official status of the proposal: (http://ec.europa.eu/prelex/detail_dossier_real.cfm?CL=en&DosId=193384 seen on September 1, 2006)

8.1.3 Restocking

Restocking has been practised by some countries for decades, generally to maintain fisheries rather than improve the stock or recruitment (Fig. 14). Restocking may be beneficial to rebuilding the stock, but it is highly unlikely that the 40% objective set by the EC will be met in all European river basins by re-stocking alone. Only a combination of several measures can be expected to bring the stock out of its current critical state. The current glass eel catches are also insufficient to re-stock inland waters, and a further decline in glass eel recruitment could result in total loss of the option to use restocking as a measure (WGEEL 2006).

8.1.4 Restoration of spawning stock

In order to restore the spawning stock, protective measures will have to be implemented. Noting the ongoing decline in the adult stock at current fishing effort (Fig. 15), also in relation to the decline in recruitment from which the current stock was derived, opportunities for protection and restoration are fading. All possible emergency measures to protect the stock from anthropogenic mortality must be implemented, the sooner the better. Beyond immediate measures, restoration plans will have to be developed and implemented, allowing the recovery of the European eel stock.

8.1.5 Long term targets and the Precautionary Approach

In accordance with the Precautionary Approach, on top of the minimum spawning stock levels an extra safety margin has been recommended. Given the many uncertainties in eel biology and management, the precautionary advice of ICES (2002) was that the European eel stock should be managed according to a precautionary target reference point of 50% of the potential maximum pristine spawner escapement. Since no further, specific information has been brought forward, the advice is continued. While the proposal of the Council regulation is for a target escapement of at least 40% of the potential biomass of adult eel, the underlying reference status of the population, in terms of silver eel biomass, is not clearly defined.

8.2 Population monitoring and control measures

Management and monitoring interconnected activities are needed. This is why the European Commission has issued a Proposal for a Community Action Plan for the Management of European Eel (COM 2003, 573), in which the international objective of restoration of the spawning stock is made explicit. The challenge for the Community is the rapid design of a management system that ensures that local measures produce results in a consistent way across the various river basins and coastal areas, Member States, and adjacent countries. To this end, criteria for sustainable management of eel fisheries will be employed, focusing primarily on recruitment of young eels to and escapement of silver eels from continental waters, and secondarily on stock abundance and anthropogenic impacts in continental waters.

Obviously, further assessment of the biological status of eel requires additional and consistent data. This is why the European eel has been included in the EU Data Collection Regulation (DCR), (Council Regulation 1543/2000 and Commission Regulations 1639/2001, 1581/2004). Required sampling levels have only been tentatively indicated, and few countries have actually included eel in their sampling programmes. The European Commission initiated a Workshop on National Data Collection for European Eel (September 2005), with the objective to specify minimum requirements on sampling levels for fishery-dependent and fishery-independent data. This report (Dekker (Ed.) 2005) presented an overview of current monitoring, surveying and sampling for eel, discussed the appropriate spatial scale for management and monitoring, develops adequate sampling intensities for sustainable management of a large number (>100) of mutually independent geographical management units, and recommended minimum requirements for future sampling in each of these management units, for each of the life stages (ibid.).

The main conclusions of this meeting were (ibid.):

1. Registration of fishing capacity, effort and landings is present in most countries, but achieves an incomplete coverage. Inland waters (of smaller size) are most frequently missed; non-commercial fisheries are substantial and almost completely unregistered.

2. Catch composition sampling occurs presently in only a few countries, but can rather easily be extended to other countries/areas. There is considerable friction between the required sampling levels (15 samples per annum per spatial management unit), the number of intended spatial management units (WFD/River Basin Districts, >100), and the size of an overall acceptable sampling programme.
3. Recruitment surveys (glass eel, young yellow eel) are in operation in most of the distribution area, but are often fishery dependent. Required coordination and harmonisation has been described before.
4. Spawner escapement surveys (silver eel) are required for evaluation of trends in the spawning stock, but not easy to implement in most areas.
5. Standing stock surveys (yellow eel) can replace silver eel surveys in unfished areas, or where silver eel monitoring is unachievable, and might provide early warning on the trends in the stock. Current practices easily allow for extensions into new areas. Coordination with and integration into WFD-monitoring is required.
6. Current monitoring data are rarely used for an assessment of the status of stock and fisheries, but the FP6-project SLIME (FP6-022488) will focus on further development of appropriate models.
7. Analysis of sampling precisions is only available in two cases; available data allow further analysis. Complications arise due to required and inherent stratification.
8. Development and implementation of national management plans will require considerable efforts. International harmonisation and exchange of methodologies can facilitate the developments.

At the WGEEL Meeting in January 2006 the monitoring objectives may be summarised as follows:

Recruitment monitoring:

It is essential that the existing recruitment indices be continued. The network of monitoring stations should be extended and strengthened to give a better coverage of spatial scale. Monitoring of glass eel gives two measures, not necessarily from the same monitoring station: firstly success of spawning escapement and oceanic larval migration and secondly, recruitment into individual catchments.

Yellow eel monitoring:

Monitoring the standing stock of yellow eel may give a useful proxy for compliance to established management targets. This may be obtained by CPUE values in the lower reaches and lakes in a catchment and where possible, the relationship between CPUE data and standing crop should be established. Together with data on size and age structure, this could provide input for modelling spawner escapement. Another approach to obtain a proxy for the standing stock is yellow eel densities (electro-fishing) in the upper parts of a catchment.

Silver eel monitoring:

Monitoring output of silver eel may be possible from mark recapture techniques. From such surveys, overall mortality in the continental phase may also be deduced. The number of case studies presently using this approach, however, is extremely limited.

8.3 Captive breeding and artificial propagation

No attempt at captive breeding of the European eel has been successful so far, and hence there is no artificial propagation.

8.4 Habitat conservation

The management measures enumerated above (8.1) to increase recruitment and spawning stock are all concerned with eel fisheries, monitoring and legal instruments. The environmental threats facing the various eel stages are discussed above (4.1). Suffice it to say that the mitigation measures that will have most positive effect on eel spawning stock are fairly long-term, and in the short time span now available for changing the trend, such measures will be too slow to prevent the eel from biological and thus commercial extinction.

8.5 Safeguards and control measures

These issues are covered above (6, 7).

9. Information on similar species, “look a like” problems and identification techniques.

As mentioned above there are around 15-17 species of so called anguillid species (genus *Anguilla*) in the world (slightly different opinions among different scientists), all of which (as far as we know) spawn in tropical waters. Of these, 14 are listed as supporting commercial fisheries and 3, subsistence fisheries and 8 species are currently utilised in commercial aquaculture (Fishbase, 2006). There are also a number of more distantly related species (eg. congers) that have generally similar morphology and in some cases could be difficult to separate, particularly in processed forms. Through genetic test *Anguilla anguilla* can be distinguished from other *Anguilla* species in all its developmental stages including glass eels. This could though be more difficult for identification of processed products and derivatives. The European eel does not overlap with any other eel species in the fishery areas of its distribution, but some overlap occurs in Icelandic rivers which also harbour American eel (*Anguilla rostrata*). The European eel has among the largest distribution of any eel species and according to FAO data bases these eels account for roughly one half of their world fishing yield but less than 10% of world aquaculture production. The Japanese eel (*Anguilla japonica*) on the other hand, with a fishing yield of roughly 10% of that of the European eel is used to produce an aquaculture output ten times the size of European aquaculture production.

In addition to traditional taxonomic characters as the specimen size and number of vertebrae, there are several new methods using DNA techniques described and applied for the identification of different *Anguilla* species, including processed products. See reference list.

10. Consultations

The current proposal has been sent to 45 range states. Responses were received from 11 countries, seven of those with comments on the proposal; Belgium, Germany, Portugal, Republic of Serbia, Slovakia, Slovenia and United Kingdom. Four countries had the intention to give further comments but did not come back. Comments was also received by some independent scientists. Relevant notifications and changes in the document have been made after the responses. The proposal was also sent to FAO.

11. Additional remarks

Assessment of the European eel under FAO's recommended criteria for CITES listing: The European eel meets the guidelines suggested by FAO for the listing of commercially exploited aquatic species. The species falls into FAO's lowest productivity category of the most vulnerable species and the rate of decline is so rapid and steep as to qualify for Appendix I listing under FAO guidelines because the eel population has declined to 20% or even less of the historic baseline (FAO 2001). FAO (ibid.) further recommend that even if a species is no longer declining, if populations (in this *the* population) have been reduced to near the extent-of-decline-guidelines, the species could be considered for Appendix II listing. The latest IUCN Red List assessment for this species is (probably) that of Sweden (2005) which lists the European eel as *Critically Endangered* (CR).

12. References

Anon. 2006. Eurostat databases, <http://fd.comext.eurostat.cec.eu.int/xtweb/>

D'Ancona, U. 1958. Comparative biology of eels in the Adriatic and the Baltic. Verh. Int. Verein. theor. Limnol. 13: 731-735.

Albert, V., Jónsson, B., Bernatchez, L. 2006. Natural hybrids in the Atlantic eels (*Anguilla anguilla*, *A. rostrata*): evidence for successful reproduction and fluctuating abundance in space and time. Molecular Ecology 15, 1903-1916.

Aoyama, J., Watanabe, S., Nishida, M. and Tsukamoto, K. 2000. Discrimination of catadromous eel species, genus *Anguilla*, using PCR-RFLP analysis of the mitochondrial 16SrRNA domain. Trans. Am. Fish. Soc. 129: 873-878.

Aoyama, J., Ishikawa, S., Otake, T., Mochioka, N., Suzuki, Y., Watanabe, S., Shinoda, A., Inoue, J., Lokman, P. M., Inagaki, T., Oya, M., Hasumoto, H., Kubokawa, K., Lee, T. W., Fricke, H. and Tsukamoto, K. 2001. Molecular approach to species identification of eels with respect to determination of the spawning site of the Japanese eel *Anguilla japonica*. Fisheries Science 67: 761-763.

Aubrun, L. 1986. Inventaire de l'exploitation de l'anguille sur le littoral de la Bretagne. Les Publications du Département d'Halieutique No 1. Ecole Nationale Supérieure Agronomique de Rennes, 124 pp.

Aubrun L. 1987. Inventaire de l'exploitation de l'anguille sur le littoral Sud-Gascogne. Les Publications du Département d'Halieutique No 5. Ecole Nationale Supérieure Agronomique de Rennes, France. 158 p.

Avise, J. C., Helfman, G. S., Saunders, N. C. and Hales, L. S. 1986. Mitochondrial DNA differentiation in North Atlantic eels: population genetic consequences of an unusual life history pattern. Proceedings National Academy of Science, USA 83: 4350-4354.

Avise, J.C., Nelson, W.S., Arnold, J., Koehn, R.K., Williams, G.C. and Thorsteinsson, V. 1990. The evolutionary genetic status of Icelandic eels. Evolution 44: 1254-1262.

Boëtius, I. and Boëtius, J. 1967. Studies on the European eel, *Anguilla anguilla* (L.). Meddelelser fra Danmarks Fiskeri- og Havundersøgelser 4: 339-405.

- Ciccotti, E., Busilacchi, S. and Cataudella, S. 2000. Eel, *Anguilla anguilla* (L.), in Italy: recruitment, fisheries and aquaculture. *Dana* 12: 7-15.
- Daverat, F., Limburg, K. E., Thibault, I., Shiao, J-C., Dodson, J. J., Caron, F., Tzeng, W-N, Iizuka, Y., and Wickström, H. 2006. Phenotypic plasticity of habitat use by three temperate eel species, *Anguilla anguilla*, *A. japonica* and *A. rostrata*. *Marine Ecology Progress Series* 308: 231-241.
- Dannewitz, J., Maes, G.E., Johansson, L., Wickström, H., Volchaert, F.A.M., and Järvi, T. 2005. Panmixia in the European eel: a matter of time ? *Proc. Royal Society of London*, 272: 1129-1137.
- Dekker, W. 1996. A length structured matrix population model, used as fish stock assessment tool. In: I. G. Cowx, Editor. *Stock Assessment in Inland Fisheries*. Fishing News Books, Oxford, England pp. 245-259.
- Dekker, W. 2000a. The fractal geometry of the European eel stock. *ICES Journal of Marine Science* 57: 109-121.
- Dekker, W. 2000b. A Procrustean assessment of the European eel stock. *ICES Journal of Marine Science* 57: 938-947.
- Dekker, W. 2003a. Status of the European eel stock and fisheries. In: Aida, K., Tsukamoto, K. and Yamauchi, K.(Eds.) *Eel Biology*, Springer-Verlag, Tokyo, 237-254.
- Dekker, W. 2003b. On the distribution of the European eel and its fisheries. *Canadian Journal of Fisheries and Aquatic Sciences* 60: 787-799.
- Dekker, W. (Ed.) 2002. Monitoring of glass eel recruitment. Report C007/02-WD, Netherlands Institute of Fisheries Research, IJmuiden, 256 pp.
- Dekker, W, Casselman, J. M, Cairns, D. K., Tsukamoto, K., Jellyman, D. and Lickers, H. 2003. Quebec Declaration of Concern: Worldwide decline of eel resources necessitates immediate action. *Fisheries* 28: 28-30.
- Dekker, W. 2004. Slipping through our hands – Population dynamics of the European eel. University of Amsterdam, the Netherlands 11/10/2004.
- Dekker, W. (Ed.) 2005. Report of the Workshop National Data Collection – European eel: Sånge Säby (Stockholm , Sweden), 6-8 September 2005.
- De Leo, G. A. and M. Gatto, 1995. A size and age-structured model of the European eel (*Anguilla anguilla* L.). *Canadian Journal of Fisheries and Aquatic Sciences* 52(7): 1351-1367.
- DeLigny, W. And E. M. Pantelouris 1973. Origin of the European eel. *Nature* 246: 518-519.
- FAO 2001. Report of the second technical consultation of the CITES criteria for listing commercially exploited aquatic species. *FAO Fisheries Report No. 667*. FAO, Rome.

Gabriel, O. 1999. Fangmethoden. *In*: Tesch F.W. (ed.), *Der Aal*. Berlin (FRG), Parey Buch Verlag, 241-288.

Gatto, M., and R. Rossi. 1979. A method for estimating mortalities and abundances of the Valli di Comacchio eels. *In* R. De Bernardi, Editor. *Proceedings of the Symposium "Biological and mathematical aspects in population dynamics"*. *Memorie dell'Istituto Italiano di Idrobiologia Dott. Marco De Marchi Suppl.* 37: 107–114.

Ginneken van, V. J. T. and Maes, G. E. 2005. The European eel (*Anguilla anguilla*, Linneaus), its lifecycle, evolution and reproduction: a literature review. *Rev Fish Biol Fisheries* 15: 367-398.

Greco, S., Melià, P., De Leo G. A., and Gatto, M. 2003. A size and age-structured demographic model of the eel (*Anguilla anguilla*) population of the Vaccarès lagoon. Internal Report 2003.47, Dipartimento di Elettronica e Informazione, Politecnico di Milano, Milano, Italy.

Holmgren, K. 1996. *On the Sex Differentiation and Growth Patterns of the European Eel, Anguilla anguilla (L.)*, Uppsala University (Acta Universitatis Upsaliensis, Uppsala). Ph D Thesis.

ICES 1999. International Council for the Exploration of the Sea. ICES cooperative research report N° 229, Report of the ICES Advisory Committee on Fisheries Management, 1998: 393-405.

ICES 2000, Report of the EIFAC/ICES working group on eels. ICES C.M. 2000/ACFM:6. (Silkeborg, Denmark), 20-24 September 1999.

ICES 2002. International Council for the Exploration of the Sea. Report of the ICES/EIFAC Working Group on Eels. ICES C.M. 2002/ACFM: 03.

ICES 2003 International Council for the Exploration of the Sea. Report of the ICES/EIFAC Working Group on Eels. ICES C.M. 2003/ACFM:06.

ICES 2004 International Council for the Exploration of the Sea. Report of the ICES/EIFAC Working Group on Eels. ICES C.M. 2004/ACFM:09.

ICES 2005a International Council for the Exploration of the Sea. Report of the ICES/EIFAC Working Group on Eels. ICES C.M. 2005/ I:01.

ICES. 2005b. Answer to Special request on Restocking of European Eel. ICES Advice

ICES 2006 International Council for the Exploration of the Sea. Report of the ICES/EIFAC Working Group on Eels. ICES C.M. 2006/ACFM:16.

Itoi, S., Nakaya, M., Kaneko, G., Kondo, H., Sezaki, K. and Watabe, S. 2005. Rapid identification of eels *Anguilla japonica* and *Anguilla anguilla* by polymerase chain reaction with single nucleotide polymorphism-based specific probes. *Fisheries Science* 71: 1356-1364. (Fast method that also works on processed material).

- Kuhlmann, H. 1975. Der Einfluss von Temperatur, Futter, Grösse und Herkunft auf die sexuelle Differenzierung von Glassaalen (*Anguilla anguilla*). Helgoländer. Wiss. Meeresunters. 27: 139-155.
- Lehman, D., Hettwer, H. and Taraschewski, H. 2000. RAPD-PCR investigations of systematic relationships among four species of eels (Teleostei: Anguillidae), particularly *Anguilla anguilla* and *A. rostrata*. Marine Biology 137: 195-204. (Used to study the systematic relationships).
- Lin, Y.-S., Poh, Y.-P., Lin, S.-M., and Tzeng, C.-S. 2002. Molecular techniques to identify freshwater eels: RFLP analyses of PCR-amplified DNA fragments and allele-specific PCR from mitochondrial DNA. Zoological studies 41: 421-430. (Four species (*A. japonica*, *A. marmorata*, *A. anguilla* and *A. rostrata*) were successfully identified with these two methods.)
- Lin, Y.-S., Tzeng, C.-S. and Hwang, J.-K. 2005. Reassessment of morphological characteristics in freshwater eel (genus *Anguilla*, Anguillidae) shows congruence with molecular estimates. Zoologica Scripta 34: 225-234.
- Limburg, K.E., H. Wickström, H. Svedäng, M. Elfman & P. Kristiansson. 2003. Do stocked freshwater eels migrate? Evidence from the Baltic suggests "Yes". Amer. Fish. Soc. Symposium 33: 275-284.
- Lintas, C., Hirano, J. And Archer, S. 1998. Genetic variation in the European eel (*Anguilla anguilla*) Molecular Marine Biology and Biotechnology 7: 263-269.
- Maes, E. G., Pujolar, J. M., Hellemans, B., Volckaert, F. A. M. 2006, a. Evidence for isolation by time in the European eel (*Anguilla anguilla* L.). Molecular ecology 15. 2095-2107.
- Maes, E.G. Pujolar, J.M., Raeymaekers, C., Joost, D. J. & Volckaert, F. 2006, b. Microsatellite conservation and Bayesian individual assignment in four *Anguilla* species. Marine Ecology Progress Series 319: 251-261.
- Moriarty, C. 1987. Factors influencing recruitment of the Atlantic species of anguillid eels. American Fisheries Society Symposium 1: 483-491.
- Moriarty C. and Dekker W. (Eds.) 1997. Management of the European Eel. Fisheries Bulletin 15, 110 pp.
- Parsons, J., Vickers, K. U. And Warden Y. 1977. Relationship between elver recruitment and changes in the sex ratio of silver eels *Anguilla anguilla* L. migrating from Lough Neagh, Northern Ireland. Journal of Fish Biology 10: 211-229.
- Pujolar, M., Maes, E.G. & Volckaert, F. (2006): Genetic patchiness among recruits in the European eel *Anguilla anguilla* Marine Ecology Progress Series 307: 209-217.
- Rehbein, H., Sotelo, C., Perez-Martin, R. I., Chapela-Garrida, M. J., Hold, G. L., Russell, V. J., Pryde, S. E., Santos, A. T., Rosa, C., Quinteiro, J. and Rey-Mendez, M. 2002. Differentiation of raw or processed eel by PCR-based techniques: restriction fragment length polymorphism analysis (RFLP) and single strand conformation analysis (SSCP). Eur. Food Res. Technol. 214: 171-177. (*A. anguilla*, *A. rostrata*, *A. japonica* and *A. australis*, all were distinguishable but in mixtures *A. anguilla* sometimes masked *A. japonica* and *A. australis*).
- Reid, K. B. 2001. The decline of American eel (*Anguilla rostrata*) in the Lake Ontario/St.

Lawrence River ecosystem: A modeling approach to identification of data gaps and research priorities. Lake Ontario Committee, Great Lakes Fishery Commission, Ann Arbor, Michigan.

Ringuet, S., Muto, F., and Raymakers, C. 2002. Eels, their harvest and trade in Europe and Asia. TRAFFIC Bullentin Vol 19 No 2.

Russel, I. C. and E. C. E Potter 2003. Implications of the precautionary approach for the management of the European eel, *Anguilla anguilla*. Fisheries Management and Ecology 10: 395-401.

Sadler, K. 1979. Effects of temperature on the growth and survival of the European eel, *Anguilla anguilla*, L. Journal of Fish Biology 15: 499-507.

Schmidt, J. 1909. On the distribution of the freshwater eels (*Anguilla*) throughout the world. I. Atlantic Ocean and adjacent region. Meddelelser fra Kommissionen for Havundersøgelser. Serie Fiskeri. 3: 1-45.

Schmidt, J. 1922. The breeding places of the eel. Philosophical Transactions Royal Society 211: 179-208.

Schmidt, J. 1925. The breeding places of the eel. Smithsonian Institute Annual Report 1924, 279-316.

Sezaki, K., Itoi, S. and Watabe, S. 2005. A simple method to distinguish two commercially valuable eel species in Japan *Anguilla japonica* and *A. anguilla* using polymerase chain reaction strategy with a species-specific primer. Fisheries Science 71: 414-421. (Same accuracy as PCR-RFLP but easier and quicker, works with *A. japonica* and *A. anguilla* in this case).

Svärdson, G. 1976. The decline of the Baltic eel population. Report of the Institute of Freshwater Research, Drottningholm 55: 136-143.

Tesch, F.-W. 1977. The Eel. Chapman and Hall, London , 434 p.

Tesch, F.-W. 2003. The eel. Blackwell Publishing, Oxford

Tsukamoto K., Nakai I. & Tesch W.-V. 1998. Do all freshwater eels migrate? Nature 396, 635–636.

Tulonen, J. and M. Pursiainen, 1992. Eel stockings in the waters of the Evo State Fisheries and Aquaculture Research Station. Suomen Kalatalous 60: 246-261.

Vøllestad, L. A. and B. Jonsson. 1988. A 13-year study of the population dynamics of the European eel *Anguilla anguilla* in a Norwegian river: Evidence for density-dependent mortality, and development of a model for predicting yield. Journal of Animal Ecology 57: 983–997.

Watanabe, S., Minegishi, Y., Yoshinaga, T., Aoyama, J. and Tsukamoto, K. 2004. A quick method for species identification of Japanese eel (*Anguilla japonica*) using real-time PCR: An onboard application for use during sampling surveys. Mar. Biotechnol. 6: 566-574. (Did discriminate *A. japonica* from two other *Anguilla* species and six other Anguilliform species, rapid method.)

Weber, M. 1986. Fishing method and seasonal occurrence of glass eels (*Anguilla anguilla* L.) in the Rio Minho, west coast of the Iberian peninsula. *Vie et Milieu*, Paris 366(4): 243-250.

Wiberg, U. 1983. Sex differentiation in the European eel (*Anguilla anguilla* L.). *Cytogenetics and Cell Genetics* 36: 589-598.

Wickström, H. 2005. *Anguilla anguilla* – ål. ArtDatabanken, SLU, faktablad, 4 p. (In Swedish)

Wickström, H. 2006. Ål (Eel). MS, 8 p. (In Swedish)

Wickström, H. and H. Westerberg 2006. The importance of eels from coastal/brackish water areas. Swedish Board of Fisheries, MS, 6 p.

Wirth, T. and L. Bernatchez 2001. Genetic evidence against panmixia in the European eel. *Nature* 409: 1037-1039.

Åström, M. 2005. Spawner escapement from yellow and silver eel fishery. Appendix 3.3 in ICES CM 2005/I:01, "Report of the ICES/EIFAC Working Group on Eels, Galway (WGEEL), 22-26 November 2004, Galway, Ireland".

Figures and Tables



Figure 1 Time trends in juvenile abundance of the major eel stocks of the world. For *Anguilla anguilla*, the average trend of the four longest data series is shown, which trend appears to occur almost continent-wide; for *A. rostrata*, data represent recruitment to Lake Ontario; for *A. japonica*, data represent landings of glass eel in Japan.

(Dekker et al. 2003)

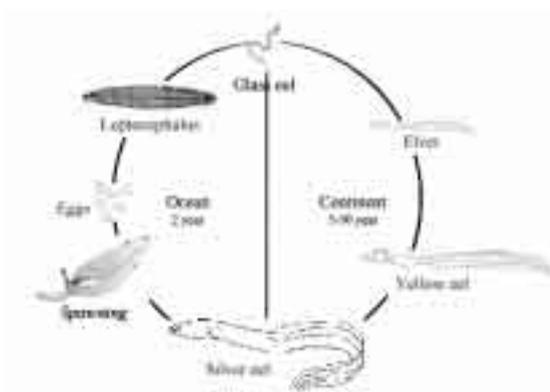


Figure 3 The life cycle of the European eel. The names of the major life stages are indicated; spawning and eggs have never been observed in the wild and are therefore only tentatively included.

(Dekker 2000a)

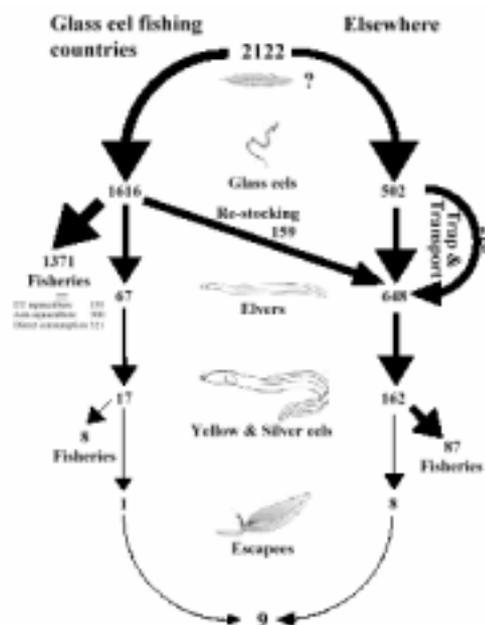


Figure 4 Dynamics of the European eel stock (numbers in millions), in the early 1990s. Estimates based on a cross-section in time, assuming a steady state. Countries with commercial glass eel exploitation to the left, other countries to the right.

(Dekker 2003a)

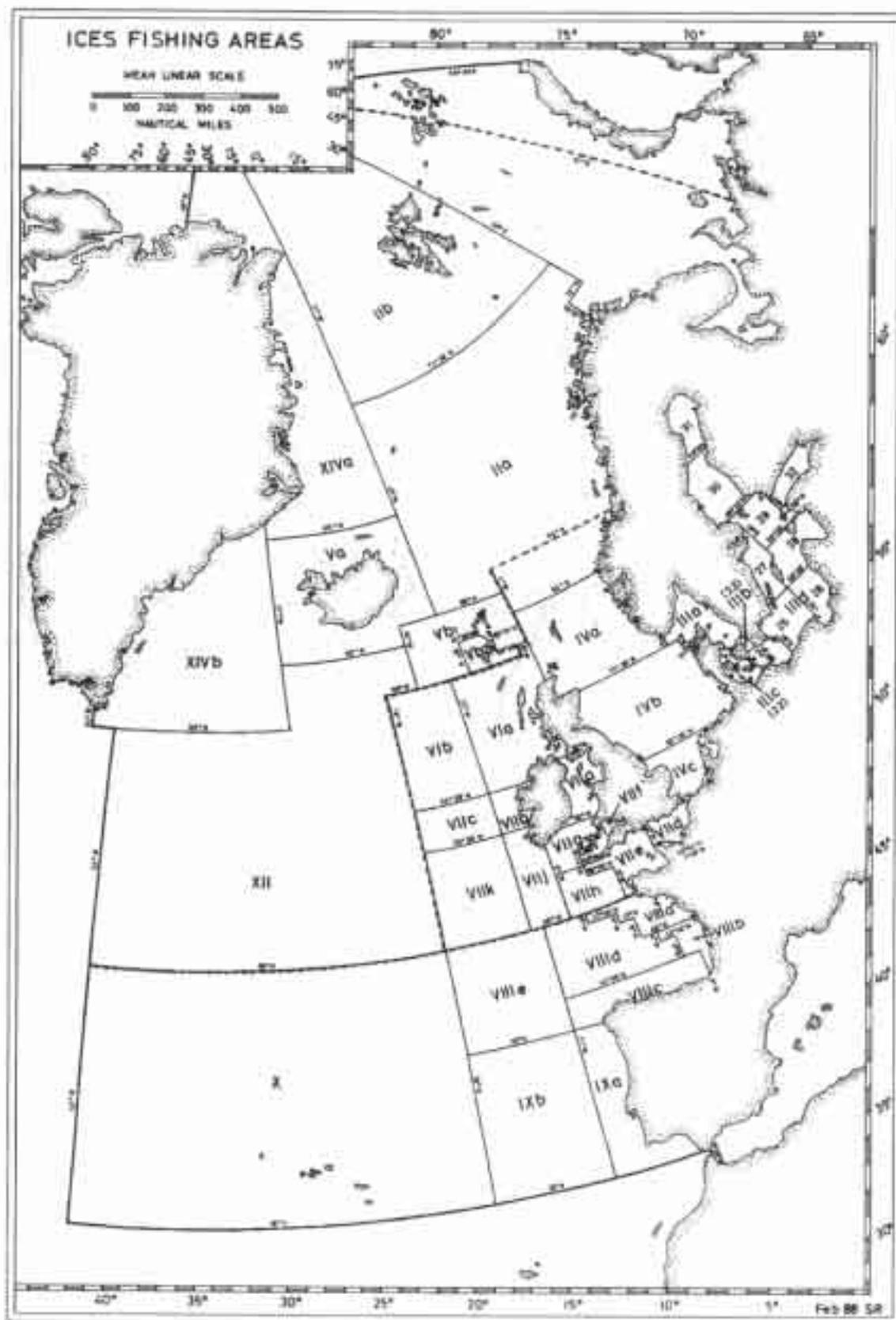
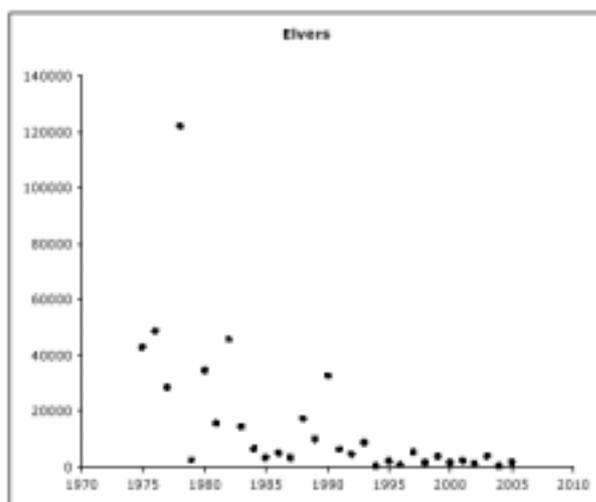
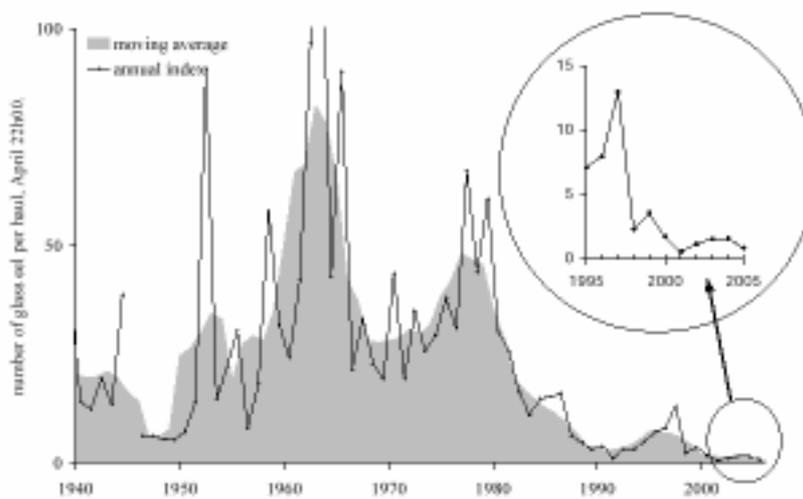
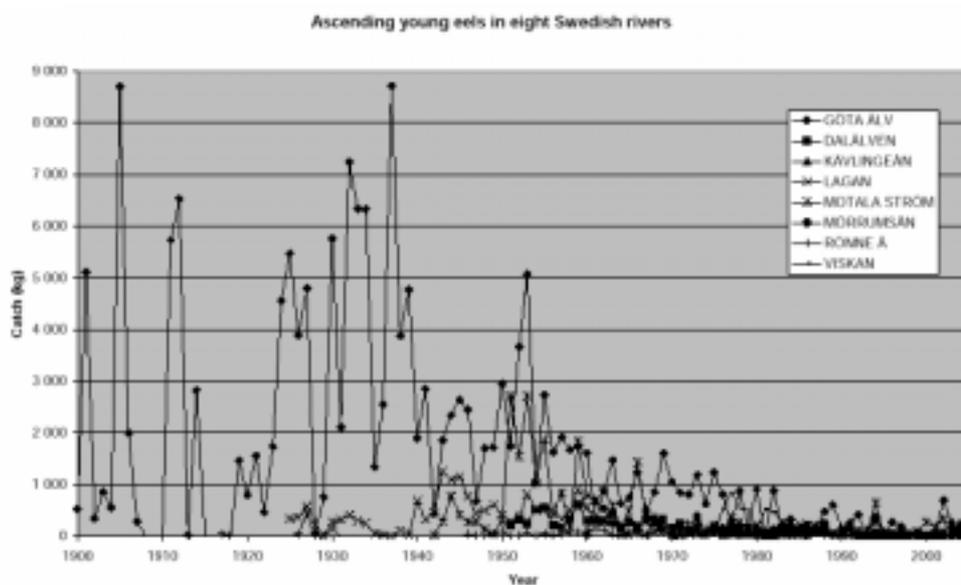


Figure 2 ICES fishing areas in the Atlantic Northeast.

Figure 5. (ICES WGEEL 2006, Annex 3: Eel stock and fisheries reported by country – 2005)



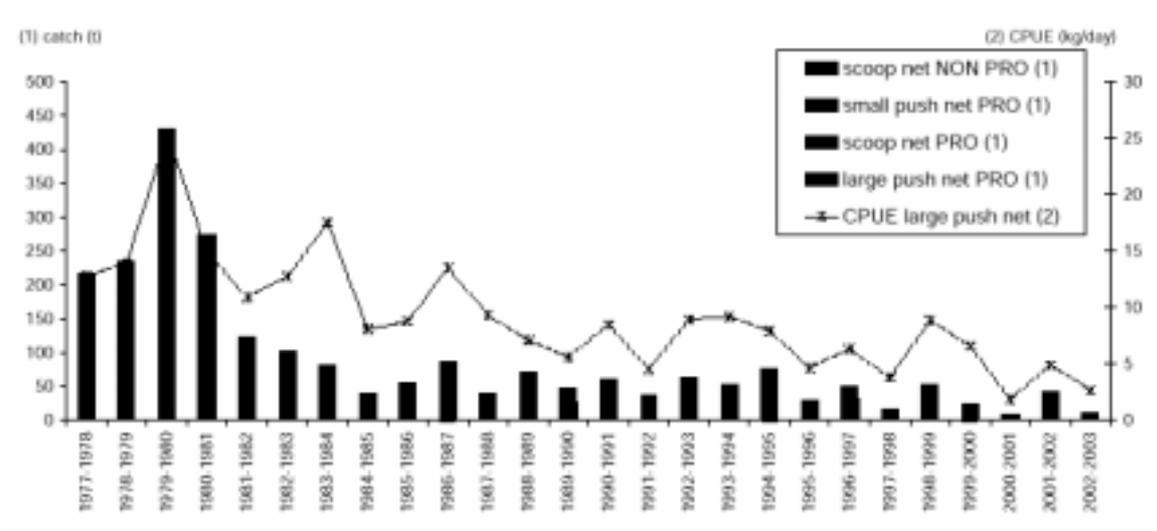
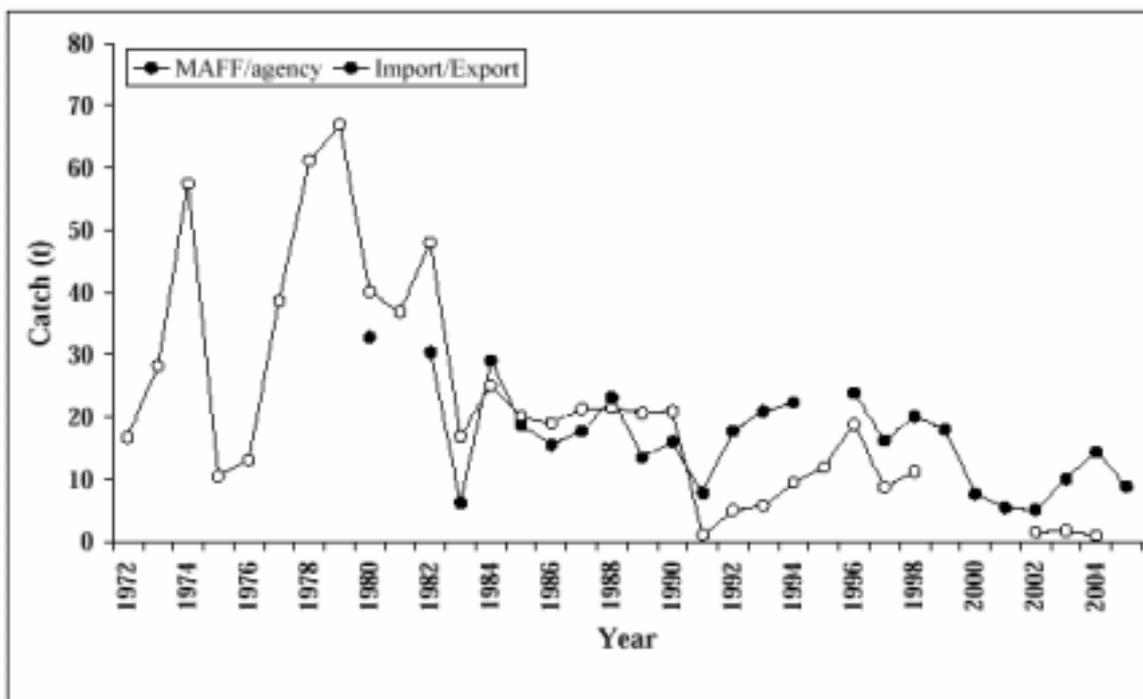
Total number of ascending elvers (number in trap) between 1975 and 2005 in the river Insa.



Time trend in the glass eel survey at Den Över.

Figure 5. continued

England & Wales annual catch of glass eel (t) from MAFF/agency data and nett export estimates (Customs & Excise), 1972 – 2005



Cumulated capture of glass eel for professional and non professional fishermen, CPUE on the Gironde basin for 1978-2003 (Source: Cemagref)

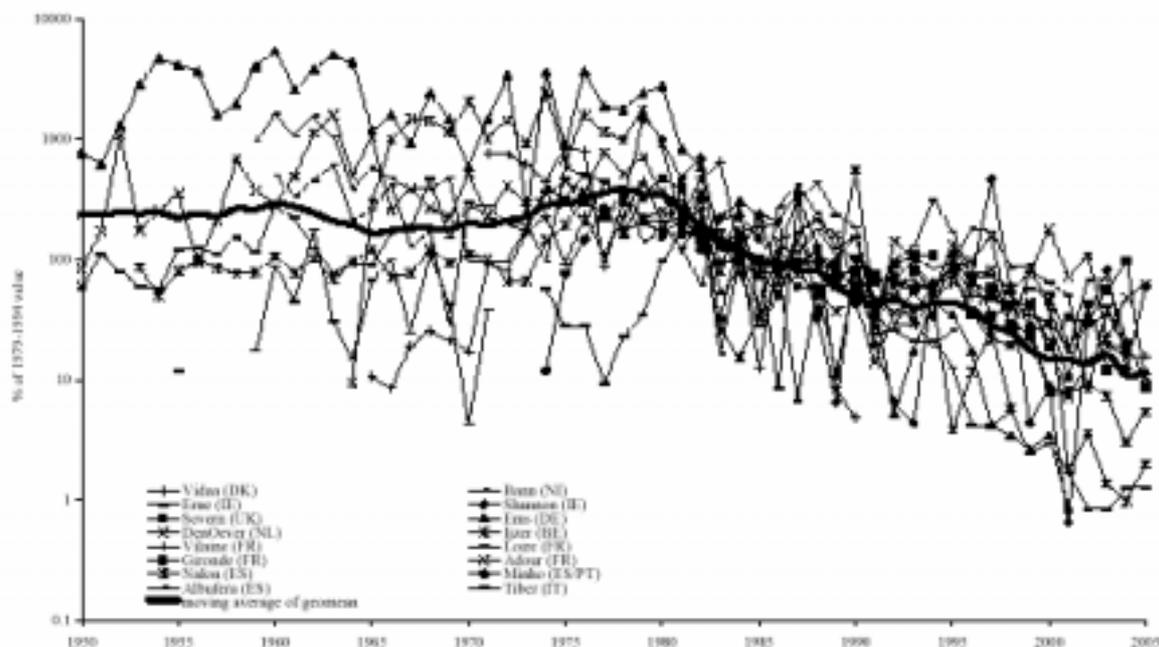


Figure 6 Time-series of monitoring glass eel recruitment in European rivers, for which data are reported for 2005. Each series has been scaled to its 1979–1994 average.

(ICES WGEEL 2005/2006)

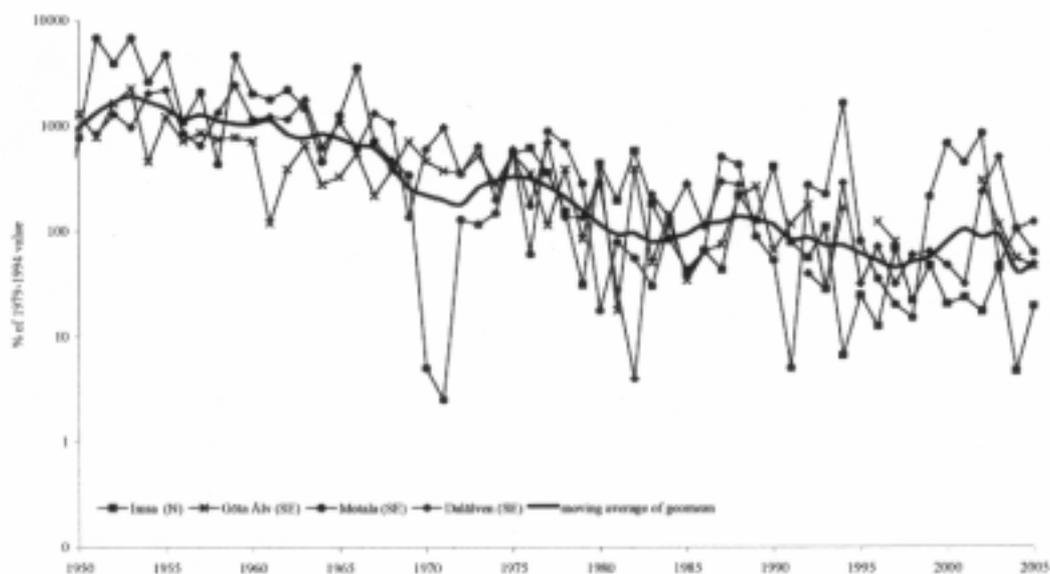


Figure 7 Time-series of monitoring yellow eel recruitment (older than one year) in European rivers, for which data are reported for 2005. Each series has been scaled to the 1979–1994 average.

(ICES WGEEL 2005/2006)



Figure 8 Re-stocking of glass eel and young yellow eel in Europe (East Germany, Netherland, Denmark, Poland, Sweden, Northern Ireland, Belgium, Finland, Estonia), in millions re-stocked. The data series of Polish re-stockings was discontinued in 1968, while the re-stockings continued.

(ICES WGEEL 2005/2006)

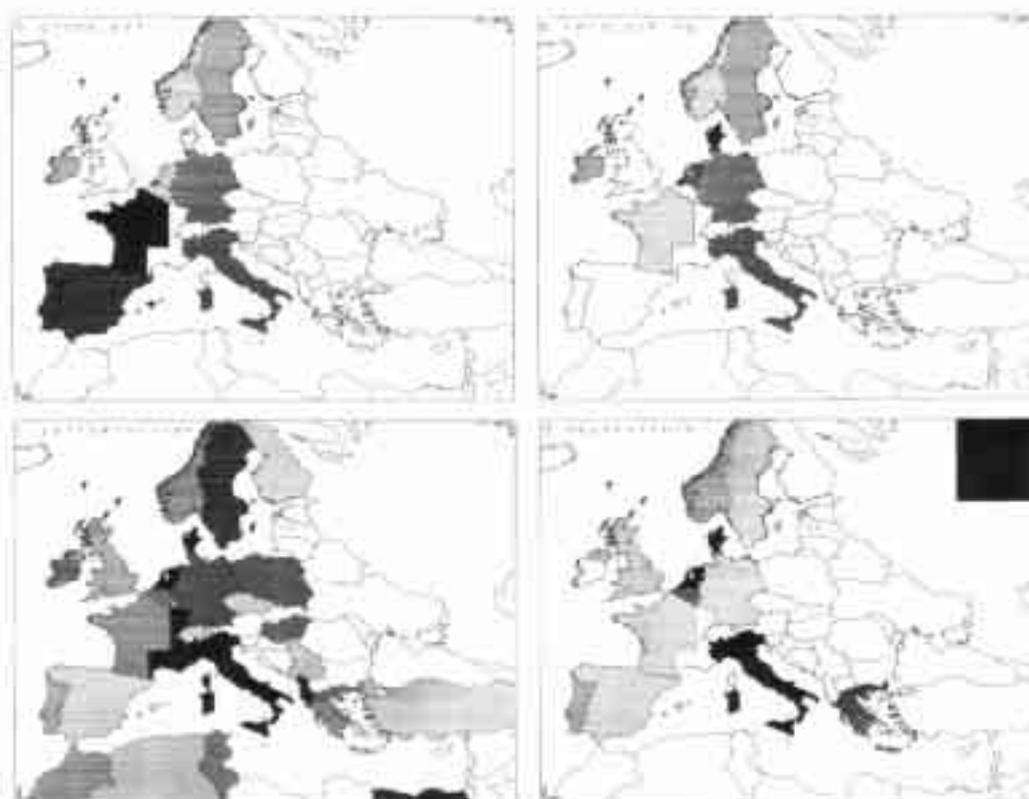


Figure 9 The spatial distribution in Europe of: a) Glass eel fisheries, b) Glass eel re-stocking, c) Yellow/silver eel fisheries and d) Aquaculture. The production of European eel in Asian aquaculture is shown in the top-right corner of panel d, in a square of equal surface area to Japan. Data from Moriarty (1997), adapted.

Legend for glass eel fisheries and re-stocking, $g \cdot km^{-2}$ land surface.



Legend for yellow and silver eel fisheries, aquaculture, $kg \cdot km^{-2}$ land surface.

(Dekker 2000b)

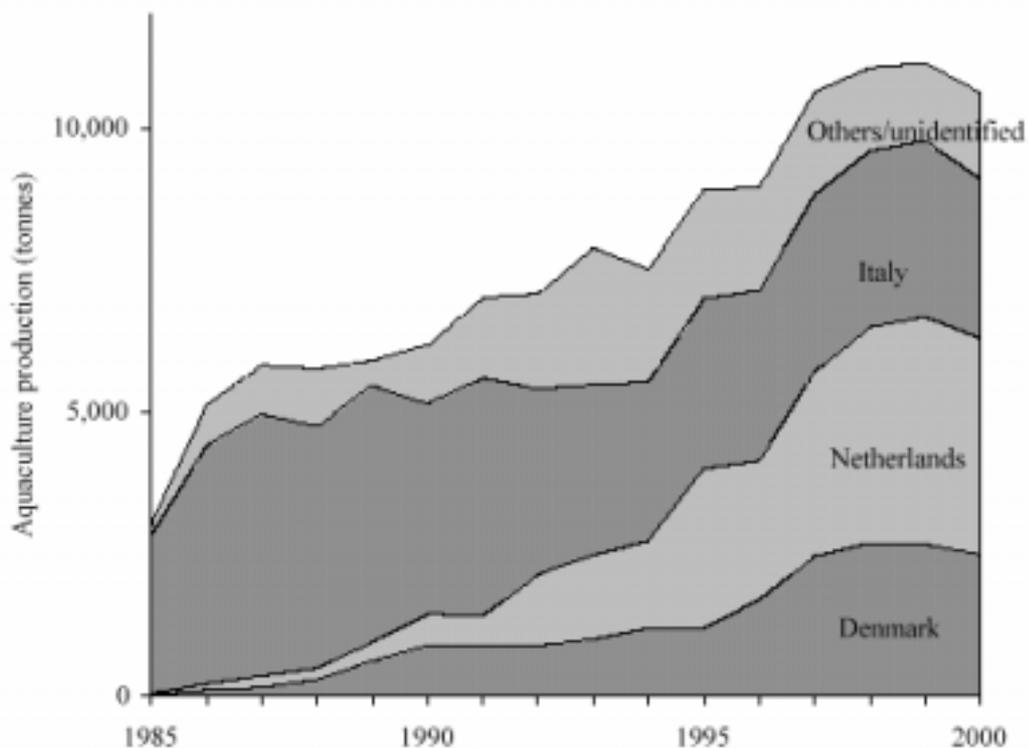


Figure10 Production of eel aquaculture in Europe. Data from ICES (2002).

(Dekker 2003a)

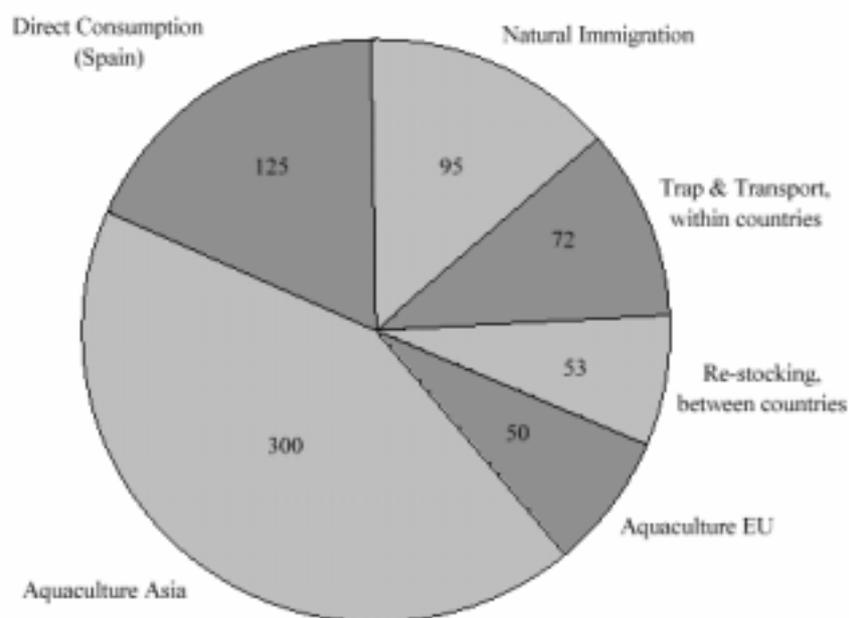


Figure 11 Disposition of glasseel landings. Numbers indicate quantities in tonnes per year. Data for the mid 1990s, from Moriarty and Dekker (1997) and (Dekker 2000).

(Dekker 2003a)



Figure 12 Catch by country and proportion of catch taken in Coastal, Estuarine and Inland water respectively. Data from UK and Ireland not divided according to catch environment.

(Wickström and Westerberg 2006)

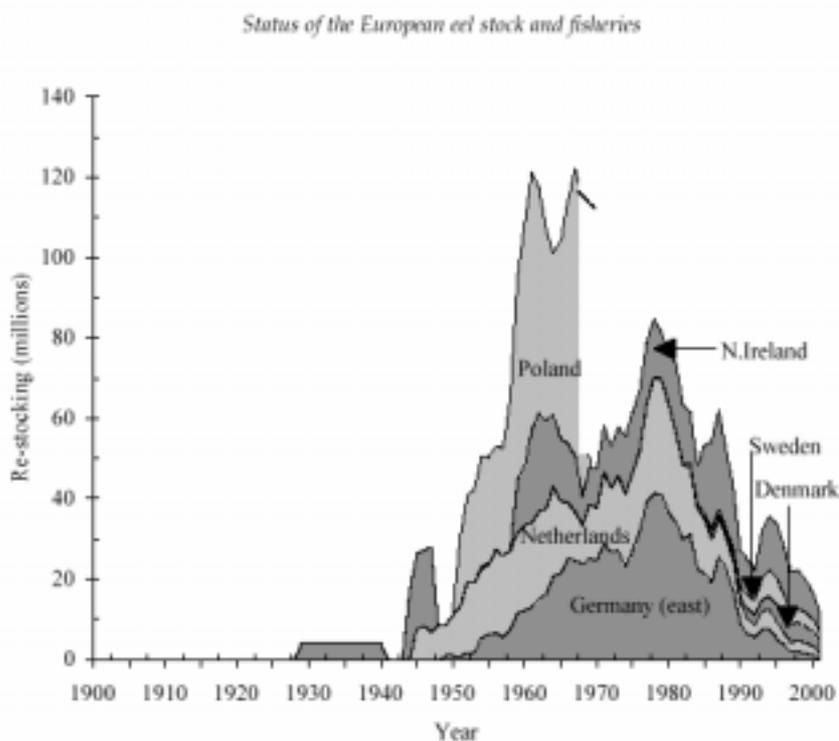


Figure 14 Re-stockings of glass eel during the 20th Century. Data from ICES (2002).

(Dekker 2003a)

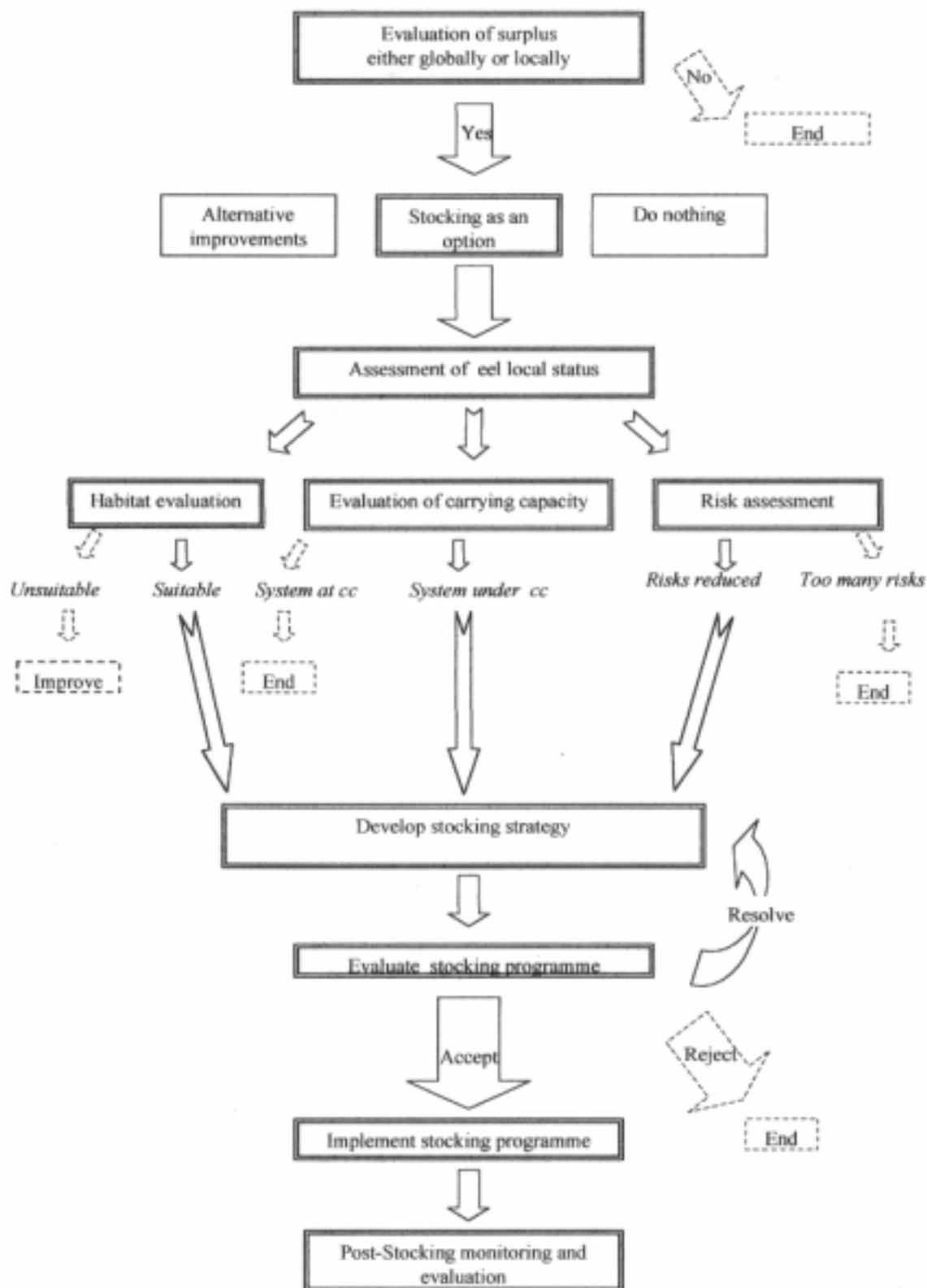


Figure 13 Stocking strategy – overview of decision process.

(ICES WGEEL 2005/2006)

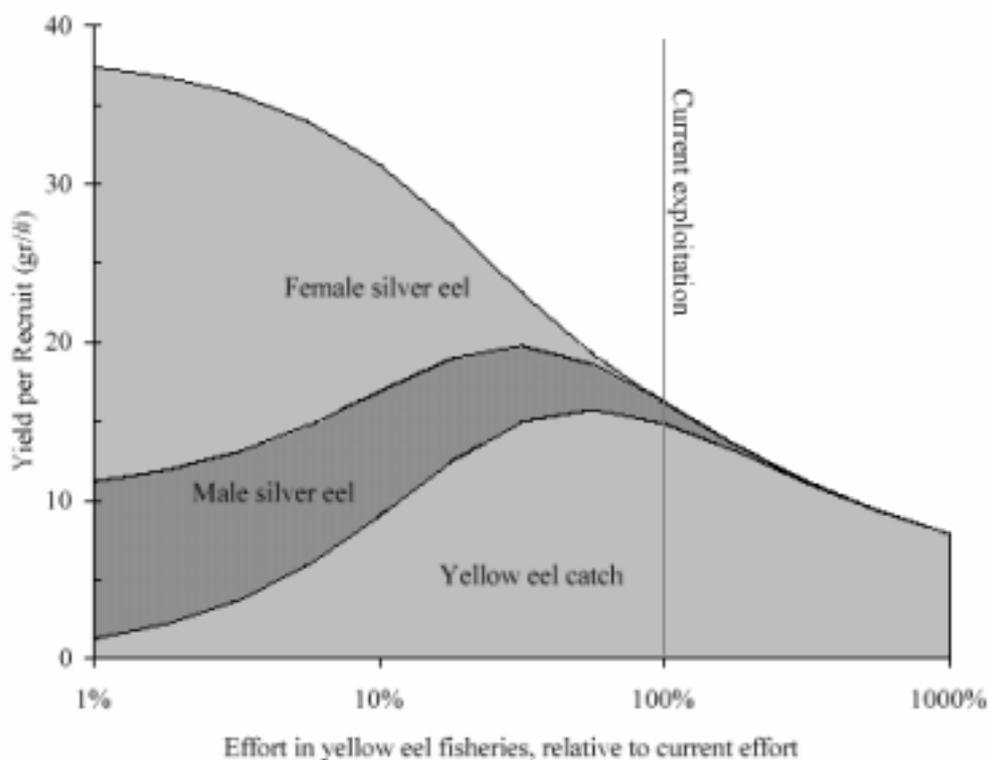


Figure 15 Yield per recruit for a mixed yellow and silver eels fishery as a function of fishing effort. Adapted from Dekker (2000).

(Dekker 2003a)

Table 1 Surface area (thousands km²) of eel habitat in freshwaters (Moriarty and Dekker 1997)

COUNTRY	EEL HABITAT (1000 KM ²)
Sweden	19
Estonia (L. Peipsi)	4
Netherland	3.4
Germany	3
Italy	2.5
Poland	2.3
Ireland	2
Great Britain	1.9
France	1.7
Spain	0.7
Denmark	0.6
Portugal	0.3
Total	41.4

(ICES WGEEL 2005/2006)

Table 2 Recruitment data series, Part I. Scandinavia and British Isles. The data units vary between data series; see the detailed Country Reports at the end of this report.

year	N Iman	S Gos Aly	S Viskan	S Motala	S Dalkhven	DK Vidua	D Ems	N.Ire. Bann	Ire. Erne	Ire. Shannon	UK Severn
1950		2947		305			835				
1951		1744		2713	210		719				
1952		3662		1544	324		1516				
1953		5071		2698	242		3275				
1954		1031		1030	309		5369				
1955		2732		1871	550		4795		167		
1956		1622		429	215		4194				
1957		1915		826	162		1829				
1958		1675		172	337		2263				
1959		1745		1837	613		4654			244	
1960		1605		799	289		6215	7400	1229		
1961		269		706	303		2995	4939	625		
1962		873		870	289		4430	6740	2469		
1963		1469		581	445		5746	9077	426		
1964		622		181.6	158		5054	3137	208		
1965		746		500	276		1383	3801	932		
1966		1232		1423	158		1840	6183	1394		
1967		493		283	332		1071	1899	345		
1968		849		184	266		2780	2525	1512		
1969		1595		135	34		1687	422	600		
1970		1046		2	150		683	3992	60		
1971		842	12	1	242	787	1684	4157	540		
1972		810	88	51	88	780	1894	2905			
1973		1170	177	46	160	641	289	2524			
1974		631	13	58.5	50	464	4129	5859	794		
1975	42945	1230	99	224	149	888	1031	4637	392		
1976	48615	798	500	24	44	828	4205	2920	394		
1977	28518	256	850	353	176	91	2172	6443	131	1.02	
1978	12181	873	533	266	34	335	2024	5034	320	1.37	
1979	2457	190	505	112	34	220	2774	2089	488	6.69	40.1
1980	34776	906	72	7	71	220	3195	2486	1352	4.5	32.8
1981	15477	40	513	31	7	226	982	3023	2346	2.15	32
1982	45750	882	380	22	1	490	674	3854	4385	3.16	30.4
1983	14500	113	308	12	56	662	92	242	728	0.6	6.2
1984	6640	325	21	48	34	123	352	1534	1121	0.5	29
1985	3412	77	200	15.2	70	13	260	557	394	1.09	18.6
1986	5145	143	151	26	28	123	89	1848	684	0.95	15.5
1987	3434	168	146	201	74	341	8	1683	2322	1.61	17.7
1988	17500	475	92	170	69	141	67	2647	3033	0.15	23.1
1989	10000	598	32	35.2		9	13	1568	1718	0.03	13.5
1990	32500	149	42	21		5	99	2293	2152	0.47	16
1991	6230	264	1	2			52	677	482	0.09	7.8
1992	4450	404	70	108	10		6	978	1371	0.03	17.7
1993	8625	64	43	89	7		20	1525	1785	0.02	20.9
1994	525	377	76	650	72		52	1249	4400	0.29	21.1
1995	1950		6	52	8		40	1403	2400	0.40	
1996	1000	277	1	14	18		20	2667	1000	0.33	14.2
1997	5500	180	8	8	8		5	2533	1038	2.12	6.6
1998	1750		5	6	15		4	1283	782	0.28	8.1
1999	3750		2	85	16		3	1345	1246	0.02	8.2
2000	1625		14	270	12		4	563	1074	0.04	3.6
2001	1875		2	178	8		1	250	699	0.00	6.4
2002	1375	685	26.2	338.8	58.6		-	1000	112	0.18	5.7
2003	3775	261	44.13	19	126.7		-	1010	580	0.38	10.8
2004	375	125	5	42	26.4		-	308	269	0.06	19
2005	1550	105	25.8	24.8	30.9				836	0.04	

(continued)

(ICES WGEEL 2005/2006)

Table 2 Recruitment data series; continued. Part 2: Mainland Europe. The data units vary between data series; see the detailed Country Reports at the end of this report.

	NE DenOever	B IJzer	F Vilaine	F Loire	F Gironde	F Gironde	F Adour	E Nalon	PK Minho	FR Tiber	Geomean ¹
				(CPUE) (Yield)							
1950	7.35			86							240
1951	14.07			166							239
1952	90.95			121							247
1953	14.78			91				14,529			243
1954	22.06			86				8,318			248
1955	30.35			181				13,576			223
1956	7.96			187				16,649			244
1957	18.2			168				14,351			230
1958	58.11			230				12,911			265
1959	31.98			174				13,071			264
1960	24.23			411				17,975			292
1961	42.05			334				13,060			278
1962	97.01			185				17,177			246
1963	138.42			116				11,507			210
1964	43.17	3.7		142				16,139			194
1965	90.39	115	5	134				20,364			168
1966	21.71	385	4	253				11,974			175
1967	33.31	575	9	258				12,977			187
1968	22.94	553.5	12	712				20,556			183
1969	19.35	443	10	225				15,628			180
1970	43.76	795	8	453				18,753			203
1971	19.53	399	44	330				17,032			194
1972	34.99	556.5	38	311				11,219			214
1973	26	356	78	292				11,056			230
1974	29.62	946	107	557				24,481	1,642		285
1975	38.05	264	44	497				32,611	10,578	11	290
1976	30.96	618	106	770				55,514	20,049	6.7	318
1977	67.32	450	52	677				37,661	36,637	5.9	360
1978	43.97	388	106	526				59,918	24,334	3.6	388
1979	60.91	675	209	642	19.7	286.2		37,468	28,435	8.4	352
1980	30.54	358	95	525.5	25.9	404.8		42,110	21,32	8.2	343
1981	26.04	74	57	302.7	20	332.2		34,645	54,208	4	263
1982	16.42	138	98	274	15	123.3		26,295	16,437	4	187
1983	10.99	10	69	259.5	13.6	80.3		21,837	30,447	4	148
1984	14.76	6	36	182.5	19.2	82		22,541	31,387	1.8	121
1985	15.3	13	41	154	9.6	64.5		12,839	20,746	2.5	97
1986	16.05	26	52.6	123.4	10.6	45.2	8	13,544	12,553	0.2	96
1987	6.25	33	41.2	145	14	82.4	9.3	23,536	8,219	7.4	83
1988	4.67	48	46.6	176.6	10.9	33	12	15,211	8,001	10.5	81
1989	3.2	30	36.7	87.1	7.2	80	9	13,574		9	55
1990	3.9	218.2	35.9	96	5.6	48.1	3.2	9,216	6	4.4	49
1991	1.18	13	15.4	35.7	7.7	64	1.5	7,117	9	0.8	42
1992	3.12	18.9	29.6	39.3	3.7	41.7	8	10,259	10	0.6	47
1993	3.14	11.8	31	90.5	8.2	69.4	5.5	9,673	7.6	0.5	40
1994	5.01	17.3	24	94.6	8.7	45.8	3	9,900	4.7	0.3	43
1995	7.12	1.5	29.7	132.5	8.2	73.2	7.5	12,500	15.2	0.3	44
1996	7.97	4.5	23.2	80.8	4.8	30.7	4.1	5,900	8.7	0.1	38
1997	12.97	9.8	22.85	70.8	6.5	50.5	4.6	3,656	7.4	0.1	29
1998	2.31	2.3	18.9	60.7	4.3	25	1.5	3,273	7.4	0.13	25
1999	3.6		16	86.9	7.5	44.1	4.3	3,815	3.8	0.06	18
2000	1.76	17.85	14.45	79.9	6.6	25.1	10	1,330	1.2	0.07	15
2001	0.58	0.7	8.46	30	1.9	9	4	1,285	1.149	0.04	15
2002	1.17	1.4	15.9	42	4.9	36.8	6	1,569		0.02	14
2003	1.56	0.539	9.37	53	2.7	10.4	1.24	1,231		0.02	16
2004	1.57	0.381	7.49	27			2.67	506		0.03	11
2005	0.85	0.787	7.36				3.5	914		0.03045	12

¹: The column **Geomean** presents the geometric mean of the three longest glass eel data series (Loire, Den Oever and Ems), after standardisation to their 1979-1994 level.

Table 3 Comparison of the 1994 and 2004 estimates of eel catches, per country. Sources: Moriarty (1997) and Moriarty and Dekker (1997); recent Country Reports at the end of this report.

COUNTRY	GLASSEEL (TON)		YELLOW + SILVER EEL (TON)	
	1994	2004	1994	2004
France	300.0	173.9	2200	1078
Italy	0.5	0.0	900	446
Spain	150.0	4.0	100	34
England and Wales	18.0	14.4	293	183
Scotland	0.0	0.0	0	0
Ireland	3.0	0.7	1035	582
Poland	0.0	0.0	1137	75
Latvia	0.0	0.0	40	12
Estonia	0.0	0.0	47	39
Sweden	0.0	0.0	1130	572
Denmark	0.0	0.0	1780	530
Norway	0.0	0.0	472	240
Belgium	0.0	0.0	0	5
Netherlands	3.0	0.0	885	920
Germany	0.0	0.0	1198	416
Portugal	20.0	4.6	0	0

(ICES WGEEL 2005/2006)

Comments to table 3 and 4: The Inland Fisheries in Portugal have commented the volumes presented for Portugal in the two tables. According to verbal comments the volume of glass eel catches are inappropriate and there should be catches of yellow and silver eel noted in table 3. Table 4 probably shows potential numbers for Portugal not the actual ones, according to the Inland Fisheries.

Table 4 Aquaculture production of European eel in Europe and Japan. Compilation of production estimates (tonnes) derived from reports of previous WG meetings, FAO, FEAP and others. Data for Sweden and the Netherlands have been revised.

	1984	1985	1986	1987	1988	1989	1990	1991	1992	1993	1994	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005	
Norway										120	200	200	200	200									
Sweden	12	41	51	90	203	166	157	141	171	169	160	139	161	189	204	222	273	200	167	170	158		
Denmark	18	40	200	240	195	430	586	866	748	782	1034	1334	1568	1913	2483	2718	2674	2000	1880	2050	1700		
Ireland																	100						
UK				20	30	0	0			25			25										
Germany											186	204	221	260	400	422	347	381	372	328			
Netherlands				100	300	200	600	900	1100	1300	1450	1540	2800	2450	3250	3500	3800	4000	4000	4200	4500	4500	
Belgium/Lux.				30	30	30	125	125	125	125	150	140	150	150	40	20	50	55					
Spain	15	20	25	37	32	57	98	105	175	134	214	249	266	270	300	425	200	259					
Portugal	60	60	900	566	501	6	270	622	505	979	200	110	200	200	200	200							
Morocco							35	41	68	85	55	55	56	42	27	28	60	28					
Algeria				72	53		22	1	0	22	20	17	17	17	22	15	18	20					
Tunisia							150	151	250	260	108	158	147	108									
Italy	2500	2400	4200	4600	4250	4500	3700	4185	3265	3000	2800	3000	3000	3100	3190	3190	2750	2500	1900	1550			
Greece		6	4	10	4	54	94	132	337	341	659	550	312	500	500	300	600	735					
Turkey																							
Macedonia								1	0	70	83	60	72	60	60	30	32						
Yugoslavia	44	52	48	49	19	10	3	1	8	2	9	5	5	5	6	6	5	4					
Croatia								7	5	5	7	6	7										
Hungary										50		50				19	19						
Czech. Rep.																							
Sum EU	1950	2229	3448	4729	5517	5159	6667	6698	6818	7721	7689	8935	9031	10646	11059	10839	10510	8435					
Japan																							10000

(ICES WGEEL 2005/2006)

Table 5.

Volume (in tonnes) of live eels (mostly glass eels) exported from the EU (Eurostat data)

(Source: Caroline Raymakers pers. comm. to CITES Scientific Authority of Sweden, 17 November 2006)

	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005	Average/year
China	8,40	23,0	106,1	40,8	45,9		17,5	56,70	36,00	30,80	47,20	41,24
Hong Kong	139,30	183,6	341,0	95,1	83,7	could not be extracted	38,1	45,60	53,50	25,80	21,40	102,71
Japan	10,70	4,2	5,5	1,1	2,6		0,9	0,10			0,30	3,18
Korea (Rep. of South)	0,80	0,0	0,0	1,2	2,8		5,9	12,90			4,10	3,46
Rest of the World	21,50	43,3	22,9	18,7	8,9		1,8	1,80	7,60	7,70	1,70	13,59
TOTAL	180,70	254,1	475,5	156,9	143,9		64,2	117,10	97,10	64,30	74,70	162,85

Table 6 and Figure 16.

EU exports of Live eels - "Anguilla spp." CN Code 0301 92 00 - Eurostat data.

(Source: Caroline Raymakers pers. comm. to CITES Scientific Authority of Sweden, 17 November 2006)

	1995	1996	1997	1998	1999	2000*	2001	2002	2003	2004	2005
tonnes	180,70	254,1	475,5	156,9	143,9		64,2	117,10	97,10	64,30	74,70
EUR/kg	72,7	99,9	133,8	202,3	154,7		253,7	175,6	194,7	388,4	705,0

*Data could not be extracted from Eurostat

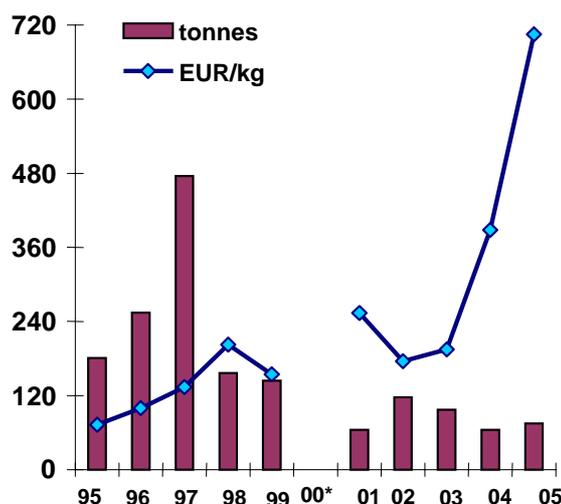


Table 7

Number (in millions) of live eels (mostly glass eels) exported from the EU (Eurostat), at 3000 glass eels per kilogramme (Han, 1999 in Ringuet *et al* 2002).

(Source: Caroline Raymakers pers. comm. to CITES Scientific Authority of Sweden, 17 November 2006)

	1995	1996	1997	1998	1999	2000	2001	2002	2003	2004	2005	Average/year
China	25,2	69,0	318,3	122,4	137,7		109,1	170,1	108,0	92,4	141,6	129,4
Hong Kong	417,9	550,8	1.023,0	285,3	251,1	could not be extracted	43,3	136,8	160,5	77,4	64,2	301,0
Japan	32,1	12,6	16,5	3,3	7,8		0,1				0,9	10,5
Korea (Rep. of South)	2,40			3,6	8,4		4,1	38,70			12,30	11,59
Rest of the World	64,5	129,9	68,7	56,1	26,7		1,4	5,4	22,8	23,1	5,1	40,4
TOTAL	542,1	762,3	1.426,5	470,7	431,7		158,0	351,3	291,3	192,9	224,1	485,1